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No. 70

THE THIRD BALTIC SEA POLLUTION LOAD COMPILATION (PLC 3)

HELSINKI COMMISSION Baltic Marine Environment Protection Commission 1998

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Preface

Within the framework of the Baltic Marine Environment Protection Commission - Helsinki Commission (HELCOM) - monitoring data on riverine and direct landbased waterborne pollution load have been collected and compiled three times since 1987. The aims are to evaluate the effectiveness of different measures taken to reduce pollution in the Baltic Sea and to determine the order of priority for different pollution sources. The first Baltic Sea Pollution Load Compilation (PLC-1) attempted to compile completely heterogeneous data submitted to the Commission on various occasions. Therefore, there were differences in the age and reliability of data, and also gaps in the data sets (Baltic Sea Environment Proceeding No 20, 1990). The Second Baltic Sea Pollution Load Compilation (PLC-2) consisted of generalised 1990 data quantifying the major pollution sources and the waterborne load with respect to BOD₇, total nitrogen and total phosphorus (Baltic Sea Environment Proceedings No 45, 1993), collected according to the PLC-2 Guidelines. However, several key data sets were missing.

This report, the Third Baltic Sea Pollution Load Compilation (PLC-3), provides an overview of the riverine and direct landbased waterborne pollution load of BOD_γ , nutrients and heavy metals to the Baltic Sea in 1995. National pollution load data were collected according to the Guidelines for the Third Baltic Sea Pollution Load Compilation (Baltic Sea Environment Proceeding No 57, 1994), but these Guidelines were not followed by all the Contracting Parties.

In this report a short description of the Baltic Sea catchment area, of the sampling, analysis and calculation methods and of the quality assurance work is given in Chapters 1 to 4. In the remaining chapters summarised information on the pollution load is presented and assessed for each Contracting Party and for nine Baltic Sea subregions showing that riverine load is the major source of pollution in the Baltic Sea. Further, a first attempt to determine how much nutrient riverine load derives from the natural background sources and how much from anthropogenic sources is presented in Chapter 6.

Some attention is drawn to the shortcomings of the data and information submitted by the Contracting Parties. The main problem was that the PLC-3 Guidelines were not fully followed by all the Contracting Parties in measuring the obligatory parameters for all sources. Consequently, it was impossible to present the total heavy metal load on the Baltic Sea.

This report concludes very clearly that the three Pollution Load Compilations can not be used for proving whether or not reduction targets (e.g. 50% reduction) have been met. The main reason is the lack of a point source and diffuse source inventory for the whole Baltic Sea catchment area permitting assessment of the anthropogenic part of the riverine input. Moreover, riverine load data are highly dependent on meteorological factors such as precipitation and runoff. It is also not possible to compare the results from PLC-3 with those of either PLC-2 or PLC-1 due to the incompleteness of the data submitted.

In accordance with the decisions of the Helsinki Commission, the Third Baltic Sea Pollution Load Compilation (PLC-3) has been carried out as a project within the "Working Group on Inputs to the Environment" (TC INPUT). Ms Heike Herata, Federal Environmental Agency, Germany, acted as Project Manager with the assistance of Mr Ain Lääne, Tallinn Technical University, Estonia.

We wish to extend sincere thanks to the representatives of all the Contracting Parties who have contributed as members of the Project Team to the success of the work not only during the expert meetings but also in the presentation of national data, the checking of results and the preparation of the report:

Mr Lars M. Svendsen, National Environmental Research Institute, Denmark, also the author of Chapter 6; Mr Enn Loigu, Tallinn Technical University, Estonia; Mr Seppo Knuuttila, Finnish

Environment Institute, Finland; Mr Horst Behrendt, Institute of Freshwater Ecology and Inland Fisheries, Germany; Ms Ilze Kirstuka and Ms Silga Strazdina, Latvian Environment Data Centre, Latvia; Ms. Judita Sukyte and Ms Aldona Margariene, Ministry of Environmental Protection, Lithuania; Ms Elzbieta Heybowicz, Institute of Meteorology and Water Management, Ms Krystyna Gazda, National Inspection Board for Environmental Protection and Mr Waldemar Jarosinski, Meteorological Institute, Poland; Mr Alexander Shekhovtsov, Center of International Projects of the State Committee of the Russian Federation on Environmental Protection, Russia; Mr Anders Widell, Swedish Environmental Protection Agency, Sweden.

In the project invaluable support has been provided by our Consultant for Quality Assurance, the author of Chapter 4, Ms Irma Mäkinen, Finnish Environment Institute and by our Consultants for Data Management Mr Antti Raike and Mr Vaino Malin, Finnish Environment Institute.

The PLC-3 work was possible only with the close co-operation of all the Contracting Parties: Denmark, Estonia, Finland, Germany, Latvia, Lithuania, Poland, Russia and Sweden, who carried out the measurements both in the rivers and at the point sources and reported the information to the data consultants.

We also wish to express our appreciation to the Finnish Environment Institute for its financial support in hosting a series of expert meetings for the PLC-3 project and for taking responsibility for the language correction and final layout of the publication.

Finally, our special thanks go to the HELCOM Secretariat for its efficient technical and financial assistance throughout the project. In particular, we wish to thank Mr. Vassili Rodionov (former Technological Secretary).

> Project Manager Heike Herata

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LIST OF ABBEREVIATIONS

Introduction

1.1 Objectives of the Pollution Load Compilations (PLCs)

According to Paragraph 1 of Article 6 of the Convention on the Protection of the Marine Environment of the Baltic Sea Area, 1974 (the Helsinki Convention), the Contracting Parties undertake to take all appropriate measures to control and minimise land-based pollution of the marine environment of the Baltic Sea Area.

In implementing the objectives of the Convention, the Helsinki Commission needs reliable data on inputs to the Baltic Sea from land-based sources in order to develop its environmental policy and to assess the effectiveness of measures taken to abate the pollution. Such data are also required for evaluation of the state of the open sea and coastal waters.

The objectives of periodic pollution load compilations (PLCs) regarding pollution of the Baltic Sea from landbased sources are:

- 1. to compile information on the direct inputs of important pollutants entering the Baltic Sea from different sources on the basis of harmonised monitoring methods;
- 2. to follow changes in the pollution load from various sources;
- 3. to determine the priority order of different sources and pollutants for the pollution of the Baltic Sea;
- 4. to assess the effectiveness of measures taken to reduce the pollution load;
- 5. to provide information for assessment of the state of the marine environment in the open sea and the coastal zones.

The task of PLC has been carried out in stages.

1.2 The three stages of the Pollution Load Compilations (PLCs)

The First Pollution Load Compilation (PLC-1)

The results of PLC-1 were published in the Baltic Sea Environment Proceedings, BSEP No. 20, in 1987. It was the first attempt to compile heterogeneous data that had been submitted to the Commission on various occasions. Because the information came from various sources there were differences in the reliability and age of the data as well as gaps in the data sets. Assuming that the values were often preliminary or based on very rough background information, it was recommended that PLC-1 should be used with caution.

The Second Pollution Load Compilation (PLC-2)

PLC-2 was implemented as a pilot programme in the measuring year 1990, aiming at basic coverage of the major aspects concerned. In order to improve the quality of the compilation, during 1988-1989 the Scientific-Technological Committee (STC) developed the Guidelines for PLC-2 that were adopted by the Commission in HELCOM Recommendation 10/4 (1989). The PLC-2 Guidelines defined the aim of the PLC and provided a harmonised methodological basis for collection and evaluation of data on a national level (for example the measuring year 1990) for evaluation of pollution source categories and parameters to be controlled. It also provided a unified methodology for measurements, calculations and reporting.

The results of PLC-2 were published in the Baltic Sea Environment Proceedings, BSEP No. 45, in 1993. The report contained the general data characterising major pollution sources and loads for nine subregions of the Baltic Sea and the Baltic Sea as a whole. The initial national information and input data were written on floppy disks, thus enabling the use of data in different model calculations.

Though the results of PLC-2 were not perfect, the second stage of the Project was a definite step forward as it

provided more reliable data, compared with the first compilation, on total loads in the Baltic Sea. Moreover, due to political changes in the Baltic Sea Region it became possible to improve reporting in the course of the project and to collect more detailed data than originally intended.

The Third Pollution Load Compilation $(PIC-3)$

PLC-3 was carried out within the ad hoc Expert Group on Pollution Load to the Baltic Sea (TC POLO). The Guidelines for PLC-3 were prepared by the lead countries - Estonia and Germany - with the assistance of experts from all Contracting Parties and are based on the recommendations of the Seminar on Monitoring of Pollution Load (14-16 April 1993, Gdansk) and the informal expert meeting on PLC-3 (15-16 June 1993, Tallinn). These Guidelines were adopted by the Commission in HEL-COM Recommendation 15/2 in 1994, and published in the Baltic Sea Environment Proceedings, BSEP No. 57, in 1994.

The Guidelines incorporated the experience gained during PLC-2 and were aimed at preparing the next Pollution Load Compilation in a way that would serve to a wider extent the purposes of the HELCOM Programme Implementation Task Force, the Technological Committee and the Environment Committee.

During the third stage of PLC the major uncertainties and weaknesses of PLC-2 could be avoided by establishing a quality assurance system and creating a data-entry system closely connected to a database. The Finnish Environment Institute (FEI), hired by HELCOM, took the lead in both. The results of the inter-laboratory comparison test were discussed during a workshop in Helsinki in October 1994 with the aim of ensuring that national laboratories be able to maintain a continuously high level of quality in routine operations. After the first version of this data-entry system for personal computers was completed, a training workshop for national data experts took place in December 1994 in Helsinki. The final version of this programme was made available to all Contracting Parties in February 1995 and was used for submission of all data compiled on a national level after the measuring period 1995.

Due to the fact that much of the pollution load is introduced into the Baltic Sea via rivers, another important step forward was to distinguish between the natural and anthropogenic contributions to riverine fluxes. After comparison of three different methods in Finland, Denmark and Germany a guide was developed in an informal expert meeting on 30-31 May 1995 in Silkeborg, Denmark.

1.3 Classification of the inputs considered in PLC-3

PLC-3 deals with discharges to the marine environment of the Baltic Sea via rivers, coastal areas and direct point source loads, with the main pollution sources as follows:

1.Riverine inputs into the Baltic Sea

- Monitored rivers
- Partially monitored rivers
- Coastal areas

2.Discharges from point sources into the Baltic Sea*

- Municipal effluents -treated
- -untreated
- Industrial effluents
- -treated
- -untreated
- Aquaculture inputs -fish farming

Airborne pollution is not dealt with in PLC-3. Information about airborne pollution load is collected by the responsible working group, TC INPUT, and published simultaneously with this report.

1.4 Parameters

reported in PLC-3

The parameters reported are classified as obligatory or voluntary according to their nature and by taking into account the detection limits of the substances in different water flows (see Table 1.1).

An overview of the entire catchment area and the nine subregions is presented in Figure 1.1. In order to make the outcome of this report comparable to the PLC-2 report, the same nine subregions of the Baltic Sea and their abbreviations were used. These are as follows:

Table 1.1 Parameters reported in PLC-3

Footnotes:

- obligarory
- voluntary
- except for rivers where BOD, and heavy metal concentrations are below the quantification limit
- ² heavy metals are obligatory for urban areas larger than 10 000 PE
- $^{\text{\tiny{3}}}$ \sim BOD $_{r}$ AOX and heavy metals are obligatory parameters for relevant industries if these parameters are regulated by sector-wise HELCOM Recommendations
- only for untreated municipal or industrial effluents
- In those cases where the recorded results are below the quantification limit, the load estimate should be supplied with the assumption that the real concentration is one-half of the quantification limit.

*Municipalities and industries discharging to the rivers downstream of the lowest water quality monitoring station should be considered as direct discharge sources. Overflows and by-passes are to be included wherever information is available.

Fugure 1. Baltic Sea catchment area and subregions

Description of the Baltic Sea catchment area

2

The total Baltic Sea catchment area comprises 1 720 270 km \neg , of which nearly 93% belongs to the Contracting Parties and 7% belongs to Non-Contracting Parties. The division of the catchment area between Contracting Parties and Non-Contracting Parties, for each of the nine Baltic Sea subregions is presented in Table 2.1. This table was compiled on the basis of information presented by the Contracting Parties (CPs) and compared with previously printed information (BSEP No. 45, Chapter 2).

The catchment areas of the Baltic Proper and the Gulf of Finland are the largest, namely 580 000 km² and 410 000 km² , respectively. The Archipelago Sea and the Sound have the smallest catchment areas. Sweden possesses the largest portion of the Baltic Sea catchment area, 440 000 km¹. The next largest territories are those of Poland, Russia and Finland, all of which possess catchment areas a bit larger than 300 000 km⁻. Germany has the smallest catchment area with 28600 km \Box . Total Baltic Sea catchment area not in the possession of Contracting Parties is 117 520 km⁻¹.

The total long-term mean flow rate via all Baltic Sea rivers is 15190 m[/s (479 km \square /a), of which nearly one-half drains into the Baltic Sea via the seven largest rivers, namely the Neva, the Vistula, the Daugava, the Nemunas, the Kemijoki, the Oder and the Göta älv. The long-term mean flow rate of these rivers and the division of the river catchment area among the different countries is presented in Table 2.2.

Much of the pollution load is introduced into the Baltic Sea via rivers. Due to the fact that the catchment areas of the rivers often belong to different countries, the pollution load discharged by several Contracting Parties also includes the load originating in other countries (both Contracting and Non-Contracting Parties) located upstream or on the other side of border rivers. The pollution load carried from the Non-Contracting Parties via the rivers is comparatively small, with the exception of that of the Nemunas where only 48% originates in Lithuania. The question of distinguishing pollution load sources among the different countries is not addressed in this report. However, a first attempt was made to estimate natural and anthropogenic contributions to riverine

fluxes (point source load and diffuse source load) for all large and many small rivers within the Contracting Parties.

For a better understanding of the load origin in different subregions, general information about population density (Map 2.1) and land use in the Baltic Sea catchment area is presented (Table 2.3 and Map 2.2). A remarkably large share (60-70%) of the territory is covered by agricultural land in Germany, Denmark and Poland (Table 2.3 and Map 2.3). The percentage of arable land in Estonia, Latvia and Lithuania is 30-50%, while the catchment area in Sweden, Finland and Russia each contain only about 10% arable land. Forests, swamps and waterbodies constitute from 65-90% of the catchment area in Finland, Russia, Sweden and Estonia. In Poland, Lithuania and Latvia they cover 30-50% of the catchment area, whereas in Denmark and Germany they cover only 19-25%. More detailed descriptions of the nine subregions of the Baltic Sea are given below in geographical order.

2.1 Bothnian Bay

The Bothnian Bay catchment area comprises 260675 km^{\Box} of which 56% (146 000 km \Box) belongs to Finland, 44% $(113 620 \text{ km})$ to Sweden and less than 1% (1 055 km \Box) to Norway. The main rivers are the Swedish Lule älv and the Finnish Kemijoki, the latter being the seventh largest river in the Baltic Sea Area.

Sweden

About 26% of the total Swedish area belongs to the Bothnian Bay catchment area. It is situated in the northern part of Sweden and is rather sparsely inhabited, with a population density of 3 inhabitants per km $(390 000)$ inhabitants in a total of 0.2% urban area). It is also heavily forested and only small areas are agricultural areas (43% forested area; 0.8% agricultural area). Furthermore, it is rich in wetlands (17%) and lakes (lake surface area, 5.9%). Other areas, including mountains, cover 33%. The length of the coastline, excluding islands, is 370 km.

The catchment area contains a large number of rivers. The main river is the Lule älv with a long-term mean flow rate of 489 m||s (1961-1990). Moreover, there are four rivers in this region, with a long-term mean flow rate exceeding 100 m \bar{N} s, for example the Torne and Kalix älven. 86% of this Swedish subregion catchment area is monitored hydrologically and hydrochemically.

Finland

About 47% of the total Finnish area is in the catchment area of the Bothnian Bay. This subregion is very sparsely populated with only 982 570 inhabitants; that means a population density of approximately 7 inhabitants per km \Box . The land is dominated by forests (61%), wetlands (29%) and lakes (5.1%) with 4.6% agriculture taking place in the southern part of the Finnish Bothnian Bay catchment area. Urban areas cover 0.3% of the land. The length of the Finnish Bothnian Bay coastline, including islands, is 4 400 km.

The catchment area contains a large number of lakes and rivers. The total river flow from this catchment area into the Baltic Sea expressed as longterm mean flow rate is 1 794 m ∇ s. The main river, the Kemijoki, has a long-term mean flow rate of 553 m $\sqrt{8}$ (1960-1995). In addition, there are three rivers with flow rates exceeding 100 m ∇ s and ten rivers have a long-term mean flow rate of between 5 and 100 m ∇ s. 92% of this Finnish subregion catchment area is monitored hydrologically and hydrochemically.

2.2 Bothnian Sea

The Bothnian Sea catchment area comprises 220 765 km \neg , of which 18% (39 300 km \Box) belongs to Finland, 80% (176 610 km \Box) to Sweden and 2% $(4 855 \text{ km})$ to Norway. The main rivers in this Bothnian Sea catchment area are the Ångermanälven and Indalsälven in Sweden and the Oulujoki in Finland.

 Table 2.2 Division of river catchment area among Contracting and Non-Contracting Parties for the seven largest rivers flowing into the Baltic Sea

[']length of the Neva to Lake Ladoga, ²length of the Göta älv to Lake Vänern, ³without delta

Sweden

About 40% of the total Swedish area belongs to the catchment area of the Bothnian Sea. It is situated in the northern part of Sweden and is sparsely inhabited with a population density of 6 inhabitants per km[]. It is also heavily forested, with small agricultural areas (1.123 million inhabitants in a total of only 0.6% urban area; 53% forested area; 1.9% agricultural area). Furthermore, it is rich in wetlands (15%) and in lakes (lake surface area: 6.4%). Other types of land areas, including mountains, cover 23%. The length of the coastline, excluding islands, is 590 km.

The catchment area contains a large number of rivers. The main river is the Ångermanälven, with a long-term mean flow rate of 494 m ∇ s (1961-1990). There are also two other rivers with long-term mean flow rates exceeding 400 m ∇ s in this subregion, one of which is the Ulme älv. In addition, there are approximately 8 rivers in this catchment area with a long-term mean flow rate above 5 m ∇ /s. 87% of this Swedish subregion catchment area is monitored hydrologically and hydrochemically.

Finland

About 14% of the total Finnish area belongs to the catchment area of the Bothnian Sea. This area has a population of 929 260, with a population density of 24 inhabitants per km . The land is dominated by forests (66.4%), wetlands (9.1%) and lakes (8.1%) . Agriculture (15.2%) is concentrated along the coast. Urban areas cover 1.2% of the land. The length of the Finnish Bothnian Sea coastline, including islands, is 6 600 km.

The catchment area contains a large number of lakes and rivers. The total river flow from this catchment area into the Baltic Sea expressed as a longterm mean flow rate is 377 m ∇ s. The flow rate of one river exceeds 100 m $\frac{N}{s}$. whereas three rivers have long-term mean flow rates between 5 and 100 m ∇ s. 85% of this Finnish subregion catchment area is monitored hydrologically and hydrochemically.

2.3 Archipelago Sea

the Aurajoki.

Finland

The catchment area of the Archipelago Sea comprises 9 000 km \Box which is completely part of Finnish territory. The main river in this subregion catchment area is

About 3% of the total Finnish area belongs to the catchment area of the Archipelago Sea. The population of this area is 458 710, with a population density of 51 inhabitants per km . The land is dominated by forests (61%), agricultural area (30%), wetlands (4.3%) and lakes (3.1%). Urban areas cover 1.7% of the land. The length of the Archipelago Sea coastline, including islands, is 20 100 km. In all coastal rivers water flow is limited. These rivers also vary greatly in flow and water quality. The total river flow from this catchment area into the Baltic Sea expressed as long-term mean flow rate is $\overline{83}$ m $\overline{1}/s$. None of the rivers have a flow rate exceeding 10 m /s and

2.4 Gulf of Finland

chemically.

four rivers have long-term mean flow rates of between 5 and 10 m ∇ s. 40% of this Finnish subregion catchment area is monitored hydrologically and hydro-

The catchment area of the Gulf of Finland comprises 412 900 km \Box of which 107 000 km \Box (26%) belongs to Finland, 276 100 km[[]] (67%) to Russia, 26 400 km[[]] $(7%)$ to Estonia and less than $0.1%$ $(3 400 \text{ km})$ to Latvia. The largest river flowing into the Baltic Sea, the Neva, is part of the Gulf of Finland catchment area and drains from the Russian territory directly into the Gulf of Finland.

A large portion of the pollution load originating in this subregion is introduced into the Baltic Sea via two big rivers, the Neva and the Narva. Because the catchment areas of both rivers belong to more than one country, the measured load also includes the load originating from all countries located upstream or on the other side of border rivers. In the case of the Neva, only 51 300 km \Box is part of Finnish territory and flows through Lake Ladoga via the river into the Gulf of Finland. However, the main part of the catchment area (215 600 km), including the river outlet, is situated in Russia. On the other hand, 39 000 km \Box (69%) of the catchment area of the Narva is located in Russia but the remaining 17200 km (31%) belongs to Estonia from which the load enters the Baltic Sea directly.

Finland

About 36% of the total Finnish area belongs to the catchment area of the Gulf of Finland. This area has a population of 2 536 330, with a population density of 24 inhabitants per km . The land is dominated by forests (64%), wetlands (10%) and lakes (17%). Agriculture (8%) takes place along the coast. Urban areas cover 1% of the land. The length of the Finnish part of the Gulf of Finland´s coastline, including islands, is 8 000 km.

The catchment area is rich in lakes, which make up almost 20% of the total catchment area. The total long-term mean flow rate from this catchment area into the Baltic Sea is 460 m /s, including one river with a flow rate exceeding 100 m
[]/s, five rivers with long-term mean flow rates of between 5 and 100 m ∇ /s and approximately ten rivers with flow rates of less than 5 m /s. 89% of this Finnish subregion catchment area is monitored hydrologically and hydrochemically.

Russia

About 1.6% of the total Russian area belongs to the catchment area of the Gulf of Finland. This area includes practically all the territory of the Saint Petersburg district, the eastern part of the Pskov district, almost all of the Novgorod district, the northwestern parts of the Tver and the Vologda districts, the western part of the Archangelsk district and the southern part of Karelia. 80% (215 600 km \mid) of the area is drained by the Neva. The total population in the Russian catchment area is 8 million, meaning a population density of 30 inhabitants per km \Box . 80% of the inhabitants live in the Saint Petersburg district. The catchment area is low and swampy. The length of the coastline, including islands, is 1 700 km.

The main rivers flow through the lakes Ladoga, Ilmen and Chudskoe (Lake Pepsi in Estonia). The retention time in Lake Ladoga is 4.5 years, in Lake Ilmen 1.5 years and in Lake Chudskoe 2.5 years. This means that a significant quantity of pollutants accumulates in these lakes. The Neva which enters the Baltic Sea directly from Russian territory has a long-term mean flow rate of

2488 m[\]/s (1859-1988). Its catchment area includes urban areas (2%), forests (55%), arable land (12%), swamps (13%), lake surface (17%) and other types of land (1%). About 70% of this Russian subregion catchment area is monitored hydrologically and hydrochemically. An additional 10% is monitored hydrologically only.

Estonia

About 60% of the Estonian territory belongs to the catchment area of the Gulf of Finland. This Estonian Baltic Sea catchment area has a population of 1.265 million, with a population density of 48 inhabitants per km^[]. On average, 30% of the catchment area consists of arable land, 39% is covered by forests and 20% by swamps. The northern section of the Estonian portion of the Gulf of Finland catchment area belongs mainly to the carst region. South Estonia is mainly part of the catchment area of Lake Pepsi (Lake Chusdskoe in Russia), discharging via the Narva into the Gulf of Finland. The sub-soil of South Estonia consists of sandstone from the devon area. The landscape is covered with small hills, lakes and bogs.

The length of the Estonian part of the Gulf of Finland´s coastline, without islands, is 600 km. The Narva, with a long-term mean flow rate of 345 m //s (1956-1982) is the principal river. About 81% of the catchment area is monitored hydrologically and 85% hydrochemically.

Latvia

Less than 5.2% of the Latvian territory belongs to the catchment area of the Gulf of Finland. The Latvian catchment area of the Gulf of Finland has a population of 47 700, with a population density of 14 inhabitants per km^[]. On average, 0.6% of the catchment area consists of urban areas, 37% is covered by forests, 45% is agricultural land and 2% lake surface. The territory is even and lowlying. In this part of the catchment area there are eight small rivers with a total long-term mean flow rate of 19.2 m $\sqrt{ }$ /s over 20-50 years of observation. 994 km (29%) of this catchment area is monitored hydrologically.

2.5 Gulf of Riga

The catchment area of the Gulf of Riga comprises 128 340 km \neg , of which 18% (23700 km) belongs to Russia, 14% $(17600 \,\mathrm{km}$] to Estonia, 39% (50 100 km]) to Latvia, 9% (11 140 km \Box) to Lithuania and 20% (25 800 km \Box) to Belarus. The main river in the Gulf of Riga catchment area is the Daugava, the fourth largest river of the Baltic Sea Area. It empties into the Baltic Sea from Latvian territory.

In this subregion the same difficulties are encountered as in the case of the Gulf of Finland concerning distinguishing the sources of pollution load via rivers to the countries. More than half of the area drained by the Latvian rivers (77 000 km \Box) is situated on the territories of Russia, Belarus, Lithuania and Estonia. Thus, the Latvian rivers serve as transit collectors for a remarkable amount of river water and, consequently, of pollution from other countries to the Baltic Sea. The most important example of this is the Daugava. Although the whole Russian subregion catchment area discharges into the Daugava, the river outlet is located in Latvia.

Estonia

About 37% of the Estonian territory belongs to the catchment area of the Gulf of Riga. The Estonian portion of the catchment area of the Gulf of Riga has a population of 295 000, with a population density of 17 inhabitants per km . About 20% of the catchment area is covered by arable land, 44% by forests and 26% by swamps. The land is low, with bogs and marshes.

The length of the Estonian part of the Gulf of Riga coastline, excluding islands, is 640 km. The main rivers are the Kasari and the Pärnu. The long-term mean flow rate of the Kasari is 25 m //s (1924 -1994) and of the Pärnu is $49.1 \,\mathrm{m}$ /s (1921-1990). About 48% of the catchment area is monitored hydrologically and 56% hydrochemically. A small portion of South Estonia discharges its river waters via the Latvian river Gauja into the Gulf of Riga.

Latvia

About 77% of the Latvian territory belongs to the catchment area of the Gulf of Riga. The population of this territory is 2 263 500, meaning a population density of 45 inhabitants per km \Box . This territory is used in the following way: urban areas 2%, forested areas 44%, agricultural areas 39% and lake surface 1.5%. The length of the Latvian segment of the Gulf of Riga coastline is 315 kilometres

The Latvian territory in this subregion is even and low-lying. About 86% of the Latvian catchment area of the Gulf of Riga is monitored hydrologically and 93% hydrochemically. The total mean flow rate in this catchment area is 1 000 m_{//s} over 20-100 years of observation. The rivers are not regulated, apart from the Daugava. The long-term mean flow rate of the Daugava is 633 m $\sqrt{8}$ (1895-1995). About 96% of the river catchment area is monitored hydrologically and hydrochemically.

Russia

About 0.14% of the total Russian area belongs to the catchment area of the Gulf of Riga which is part of the catchment area of the River Daugava (the Sapadnaja Dvina in Russia) and has no direct outlet to the open sea. It has a population of 150 000, meaning a population density of 6 inhabitants per km[]. This area is situated west of the Valday Uplands. The main river is the Daugava which, along with seven of its tributaries originates here. The largest two tributaries are the Meza and the Lutshessa. The land is low and swampy without any large industrial centres or cities. Forests and agricultural areas dominate.

Lithuania

About 17% of the Lithuanian territory belongs to the catchment area of the Gulf of Riga (through the rivers Musa (the Meza in Russia), Birvyte and Laukesa (the Lutshessa in Russia)), which drains entirely via Latvian territory into the Gulf of Riga. Most of the area is monitored hydrologically and hydrochemically. The territory has a population of 313 600, with a population density of 26.5 inhab-itants per kmⁿ. The Lithuanian sub-region catchment area is dominated by agriculture (53.6%) and forests (31.3%), with 4.8% urban areas, 4.1% water bodies, 2.4% wetlands and 3.8% de-voted to various other uses.

2.6 The Baltic Proper

The catchment area of the Baltic Proper comprises 574 245 km⁻, to which all Contracting Parties except Finland belong, as well as Non-Contracting Parties Belarus, the Czech Republic, Ukraine and Slovakia with totally 78 360 km \Box (14%). The catchment areas of the Contracting Parties are divided as follows: 3% (15 000 km[]) belongs to Russia, 0.2% (1 100 km \Box) to Estonia, 2% (11 100 km\]) to Latvia, 9% (54 160 km^[]) to Lithuania, 54% (311 900 km^[]) to Poland, 2.6% (18 200 km $[]$) to Germany, 0.2% (1 200 km \Box) to Denmark and 15% $(83 225 \text{ km})$ to Sweden.

Three of the seven largest rivers are situated in the Baltic Proper catchment area. Two of them, the Vistula and the Oder, enter the Baltic Sea from Polish territory. The third biggest river, the Nemunas, flows from the Lithuanian territory through the Curonian Lagoon into the Baltic Sea. In this subregion there are also many smaller rivers situated in the different countries. Therefore, the measured river pollution load also includes loads originating in all other countries located upstream or on the other side of the border rivers.

The total catchment area of the Vistula comprises 194 420 kmⁿ, of which 87%, populated by 22.3 million inhabitants, belongs to Poland. 12 600 km belongs to Belarus, 11 170 km \Box belongs to Ukraine and 1950 km⁻ belongs to Slovakia. The total catchment area of the Oder comprises 118 840 km¹. 89% of this catchment area, populated by about 13.1 million inhabitants, belongs to Poland. The catchment area of the Oder also includes 6% of the Czech Republic (1.4 million inhabitants) and 5% of Germany (0.4 million inhabitants). Another 10 406 km $\sqrt{\ }$ of the Polish territory, populated by nearly 1 million inhabitants, is within the catchment areas of the Pregel and the Nemunas and of smaller rivers, flowing into the Baltic Sea via Russia and Lithuania.

The Nemunas, which discharges into the Baltic Proper from the Lithuanian territory (46 700 km $[]$), drains areas in Belarus (45 450 km⁻]), Poland (2 510 km^[]), Russia (3 170 km^[]) and Latvia (90 km \Box). On the other hand, 7 459 km \Box of the Lithuanian territory belongs to the catchment areas of the River Venta and the River Bartuva, flowing into this Baltic Sea subregion through the Latvian territory. The River Sventoji drains directly to the Baltic Sea.

Estonia

About 3% of the Estonian territory, namely the western parts of the Islands Saaremaa and Hiiumaa, belongs to the catchment area of the Baltic Proper. That portion of the catchment area has a population of 10 000, with a population density of 9 inhabitants per km . The length of Estonia´s Baltic Proper coastline, excluding small islands, is 570 km. The territory consists of 14% arable land, 55% forests and 25% swamps. In it, there are neither rivers nor direct pollution sources related to the PLC-3 monitoring programme.

Latvia

Nearly 17% of the Latvian territory belongs to the catchment area of the Baltic Proper. The territory has a population of 311 500, with a population density of 28 inhabitants per km . It consists of 1.5% urban areas, 48% forests, 38% agricultural areas and 0.7% lake surface. The length of Latvia´s Baltic Proper coastline is 189 kilometres. The Latvian Baltic Proper catchment area is even and low-lying. About 63% of the catchment area is monitored hydrologically and 66% hydrochemically. Its total mean flow rate is 135 m ∇ /s over 20-60 years of observation. The main rivers are the Venta, the Barta (the Bartuva in Lithuania) and the Saka.

Lithuania

Nearly 83% of the Lithuanian territory belongs to the catchment area of the Baltic Proper, including the river catchment areas of the Nemunas, the Bartuva, the Venta and the Akmena-Dane. The population of this territory is 3 404 400, meaning a population density of 57 inhabitants per km^[]. The Lithuanian subregion catchment area is dominated by agriculture (54%) and forest (31%), with 5% urban areas, 4% water bodies, 2% wetlands and 4% devoted to various other usages.

The main river, the Nemunas, discharges into the semiclosed Curonian Lagoon. The retention time for Nemunas discharges in the Curonian Lagoon in the case of full mixing is four months. The length of the Lithuanian part of the Baltic Proper coastline, including the Curonian Lagoon, is 99 kilometres. The long-term mean flow rate of the Nemunas is 664 m \neg /s (1811-1995). About 96% of the catchment area of the Nemunas is monitored hydrologically and about 95% hydrochemically.

Russia

About 0.1% of the Russian territory belongs to the catchment area of the Baltic Proper, namely the Kaliningrad region. The main rivers are the Pregel and the Nemunas. The total population is 878 000, meaning a population density of 58 inhabitants per km¹. The total catchment area is monitored hydrologically and hydrochemically. The length of the Russian part of the Baltic Proper coastline, including islands, is 200 km. The largest part of the catchment areas of the rivers Pregel and Nemunas are situated in Belarus and Lithuania.

Poland

Almost all of the Polish territory (99.7%) belongs to the catchment area of the Baltic Proper. This area has a population of over 38 million, with a population density of 123 inhabitants per km \neg . 62% of the population is concentrated in urban areas. The remainder lives in agricultural regions, constituting 60% of the territory, (46% of which is arable land,

1% of which is orchard and 13% of which is grassland). The entire catchment area consists of 29% forest, 3% waterbodies and 6% urban areas. The usage of the remaining 2% is not specified. The length of the Polish coastline, including that of the Hel Peninsula and the islands on the Baltic Sea side, is 528 km.

The main rivers in the Polish part of the Baltic Proper catchment area are the Vistula, which flows into the Gulf of Gdansk, and the Oder, which flows into the Pomeranian Bay through the Szczecin Lagoon. The Polish Vistula catchment area comprises $168,700$ km with a long-term mean flow rate of 1 081 m| /s (1951-1990). The Polish Oder catchment area comprises 106 060 km with a long-term mean flow rate of 574 m³/s (1951-1990). About 99.7% of the Polish catchment area of the Baltic Proper are monitored hydrologically and hydrochemically.

About 35% of the monitored river water and 40% of waste water flows through lagoons and coastal lakes before entering the sea. These reservoirs, with retention times of several weeks, are affected by periodic inflows of sea water. Therefore, pollution load monitoring in the outflow into the sea is very difficult. The processes of degradation and of pollution accumulation that take place during the long retention time in the reservoirs also cause a significant decrease in pollution load in comparison with the monitored load.

Germany

Nearly 4% of the German territory belongs to the catchment area of the Baltic Proper. This comprises the main part of the area of Mecklenburg-Western Pomerania as well as the Oder catchment basin of the Federal States of Brandenburg and Sachsen. Approximately 1.56 million people live in the German Baltic Proper catchment area, meaning a population density of 86 inhabitants per kmⁿ. Stralsund, Greifswald and Neubrandenburg are the population centres of this region. Land use is devided between agriculture, forestry and food production. About 70% of these combined areas are fields and grasslands, 17% is covered by forests and nearly 4% by water.

The length of the coastline along the open sea is 161 km, whereas the "bodden coastline is 1 135 km. Bodden is a specific term in Germany for shallow bays separated by spits of land or islands and peninsulas along the coast. Bodden coastline is typical of Mecklenburg-Western Pomerania. Because of changing water levels and currents and the effect of the surf, the coastline is always changing. The open sea coastline is particularly affected: 70% of it recedes 0.2-0.4 m per year.

The main rivers in the Baltic Proper area are the Peene and the Uecker. The catchment area of the Peene comprises 5110 km , and the long-term mean flow rate is 23.6 m[$\frac{1}{5}$ (1977-1994). The Uecker has a catchment area of 2 401 km[[]] and a long-term mean flow rate of 7.8 m[/s (1977-1994).

Denmark

Nearly 3% of the Danish territory, consisting of the islands of Zealand, Falster and Bornholm, belongs to the catchment area of the Baltic Proper. This area has a population of 82 400, meaning a population density of 68 inhabitants per km \neg . 65% of the Danish Baltic Proper catchment area consists of arable land, 62% of which has been used for cereal cultivation. Forests cover about 22%, while meadows, moorlands and lakes cover about 2%. In total, natural and cultivated areas cover nearly 89% of the land. The length of the coastline in this subregion is nearly 443 km. Only 28% of the Danish Baltic Proper catchment area is monitored using the streams. The total long-term mean flow of Danish rivers into the Baltic Proper is 1.95 m ∇ s (1971-1990). The main river is the Mern with a long-term mean flow rate of 0.41 m_{$\sqrt{$ s (1971-1990).}

Sweden

Nearly 19% of the Swedish territory belongs to the catchment area of the Baltic Proper, which is heavily forested (52%), but is also more densely populated than catchment areas further north. 4.1 million people live there in a total of 2.6% urban area; meaning a population density of 48 inhabitants per km \neg . The agricultural area is larger than in the

north, covering 16% of the catchment area. Wetlands and lake surface cover 3% and 10% of the land, respectively. Other types of terrain, including mountains, cover 16%. The length of the coastline, excluding islands, is 1 190 km. The major river is the Norrström, the outlet of Lake Mälaren through Stockholm, which has a long-term mean flow rate of 166 m||/s (1961-1990). Moreover, there are approximately ten rivers in the catchment area with a long-term mean flow rate above 5 m //s. Approximately 68% of the Swedish catchment area is monitored.

2.7 Western Baltic

The catchment area of the Western Baltic comprises 22 740 km \Box , of which 46% $(10 400 km$]) belongs to Germany and 54% (12 340 km \Box) to Denmark. There are no big rivers. Most of the pollution load comes into the marine environment via many small rivers each of which has a long-term mean flow rate of less than 20 m /s.

Germany

About 3% of the German territory belongs to the catchment area of the Western Baltic. The eastern third of the Federal State of Schleswig-Holstein and the western part of the Federal State of Mecklenburg-Western Pomerania are located in this subregion catchment area. The total population of the runoff area is approximately 1.74 million inhabitants (1.1 million in Schleswig-Holstein and 0.64 million in Mecklenburg-Western Pomerania), with a population density of 159 per km^[]. The main centres of population in Mecklenburg-Western Pomerania are Rostock (230 000 inhabit-ants) and Wismar (51 000 inhabitants). In Schleswig-Holstein 50% of the total population lives in cities with more than 80 000 inhabitants. The largest populations and industrial centres are Kiel, Lübeck, Flensburg and Schleswig.

The catchment area in Schleswig-Holstein consists of 9% forests, 6% urban areas, 5% inland waters and nearly 80% agricultural land. The total length of the coastline is 521.5 km of which 193.5 km is open sea coastline belonging to Mecklenburg-Western Pomerania and 328 km is situated in Schleswig-Holstein.

The catchment area is a postglacial moraine landscape. It drains into the southern part of the highly structured Western Baltic, which includes subbasins known as Bay of Mecklenburg, Bay of Wismar, Bay of Lübeck, the Kiel Bight and the Fehrman Belt. With sandy marl as the main soil material, the following other types of soil prevail in the catchment area: stagnic or other gleysoils, cambisoils and agrisoils. Humic gleysoils and fluvisoils are found in lowlands and along watercourses.

The main river in Mecklenburg-Western Pomerania is the Warnow with a catchment area of $2982 \,\mathrm{km}$ and a longterm mean flow rate of 17.1 m ∇ s (1974-1994). There are two big rivers in Schleswig-Holstein: the Trave with a catchment area of 1807 km and a longterm mean flow of 7.5 m $\sqrt{ }$ (1971-1992), and the Schwentine with a catchment area of 714 km \Box and a long-term mean flow rate of 4.3 m[/s (1971-1992).

Denmark

Nearly 29% of the Danish territory, with a population of 1.6 million, belongs to the catchment area of the Western Baltic. Population density in this area is approximately 128 inhabitants per km \Box . The second and third largest Danish towns discharge into the Western Baltic. The Danish Western Baltic catchment area includes 68% arable land, of which 62% has been used for cereal cultivation. Forests cover about 14%, while meadows, moorlands and lakes, cover about 3%. Thus, natural and cultivated areas cover nearly 87% of the land. The remainder consists of consolidated areas: roads, villages and towns. The length of the coastline in this subregion is nearly 3 650 km. The area is covered mainly by Pleistocene fluvio-glacial sedimentary deposits with loam, sandy loams and loamy sand as dominant soil types. The elevation is low and slopes steeper than 6% only occur in about 2% of the total land mass.

More than 48% of the Danish Western Baltic catchment area is intensively monitored via numerous stations in the streams. The long-term mean flow rate from these Danish rivers into the marine areas is $50 \text{ m} / \text{s}$ (1971-1990) for an area-specific runoff of about 267 mm. None of the seven largest Danish rivers flowing into the Western Baltic has a long-term mean flow rate exceeding 20 m \sqrt{s} ; for example the Suså has a flow rate of only 6.8 m ∇ s, the Vejle 6.6 m $\sqrt{ }$ s and the Odense 6.5 m $\sqrt{ }$ s.

2.8 The Sound

The catchment area of the Sound comprises 4625 km , of which nearly 38% (1740 km) belongs to Denmark and 62% (2 885 km \Box) to Sweden. The main rivers in the Sound are the Tryggevalde in Denmark and the Kävlingeån in Sweden.

Denmark

Nearly 4% of the Danish territory with 1.5 million inhabitants belongs to the catchment area of the Sound. The population density of this region is 849 inhabitants per km . This catchment area of the Sound consists of about 43% arable land, 58% of which has been used for cereal cultivation. Forests cover about 18%, while meadows, moorlands and lakes, cover about 5%. All natural and cultivated areas cover nearly 66% of the land. The length of the coastline in this subregion is nearly 429 km.

Approximately 64% of the Danish Sound catchment area is monitored via the streams. The total mean flow rate from these Danish rivers to the marine areas is 6.1 m $\frac{1}{s}$ (1971-1990), equivalent to an area-specific runoff of about 175 mm. The main river is the Tryggevalde, with a long-term mean flow rate of 2.2 m $\sqrt{ }$ s (1971-1990).

Sweden

Approximately 0.6% of the Swedish territory belongs to the catchment area of the Sound. This catchment area is

clearly different from all other Swedish catchment areas in that it contains a large share of agricultural land (64%). It also differs in population density, as there are no less than 625 000 inhabitants in this small area, meaning a population density of 240 inhabitants per km^[]. Urban areas cover 6% of the land. Small areas are covered by forests (10%), wetlands (0.7%) and lakes (1.3%). Other types of terrain, including mountains, cover 18%. The length of the coastline, excluding islands, is 80 km. Five rivers have a mean flow rate of above 2 m //s, for example the Saxån and the Segeå. The major river is the Kävlingeån, with a long-term mean flow rate of 12 m $\sqrt{ }$ s (1961-1990). About 90% of the Swedish catchment area is monitored.

2.9 The Kattegat

The catchment area of the Kattegat comprises 86 980 km⁻, of which 18% (15 830 km \Box) belongs to Denmark, 73% (63 700 km \Box) to Sweden and 9% (7 450 km \Box) to Norway. The main river is the Göta älv in Sweden, which is the seventh largest river flowing into the Baltic Sea.

Denmark

About 37% of the Danish territory with 1.5 million inhabitants belongs to the catchment area of the Kattegat. Population density in this region is 92 inhabitants per km . The catchment area consists of 66% arable land, of which 54% has been used for cereal cultivation. Forests cover about 16%, while meadows, moorlands and lakes, cover about 5.5%. In all, natural and cultivated areas cover nearly 88% of the land. The remaining part consists of consolidated areas: roads, villages and towns. The length of the coastline in this subregion, including islands, is nearly 2 500 km. The area is covered mainly by Pleistocene fluvio-glacial sedimentary deposits. The elevation is low and slopes steeper than 6% occur in only about 2% of the total land mass. Sandy soils dominate in western and northern Jutland.

More than 61% of the Danish Kattegat catchment area is intensively monitored via numerous stations in the streams. The total long-term mean flow rate from these Danish rivers into the marine areas is 156 m ∇/a (1971-1990), equivalent to an area-specific runoff of about 311 mm. There is one large river, the Gudenå, discharging into the Kattegat with long-term mean flow rate of 32.5 m /s (1971-1990). The second and third largest rivers draining into the Kattegat are the Karup, with a long-term mean flow rate of 9.5 m $\frac{1}{s}$ (1971-1990) and the Skals, with a long-term mean flow rate of 5.0 m $\sqrt{ }$ s (1971-1990).

Sweden

About 14% of the Swedish territory belongs to the catchment area of the Kattegat. Except for its size this catchment area is basically similar to the Swedish portion of the Baltic Proper catchment area. Thus, it consists of 1.8% urban area and has 2.136 million inhabit-ants for a population density of 30 inhabitants per km⁷. Forests cover 45% of the land and 12% is used for agricult-ure. Wetlands and lakes cover 7.3% and 14.2% of the land, respectively. Other terrains, including mountains, covers 20%. The length of the coastline, exclud-ing islands, is 250 km. The major river is the Göta älv with a long-term mean flow rate of 572 m s (1961-1990). Approx-imately five other rivers have a long-term mean flow rate exceeding 20 m ∇ s; for example the Lagan, the Nissam and the Åtran. About 90% of the Swedish catchment area is monitored.

Map production by Beijer Institute of Ecological Economics, UNEP, GRID Arendal and Systems Ecology, Stockholm University.

Map 2.1. Population density classes of the Baltic Sea catchment area

Map production by Beijer Institute of Ecological Economics, UNEP, GRID Arendal and Systems Ecology, Stockholm University.

Map 2.2. Land cover classes of the Baltic Sea catchment area

Map production by Beijer Institute of Ecological Economics, UNEP, GRID Arendal and Systems
Ecology, Stockholm University.

Map 2.3 Percentage of arable land area in the Baltic Sea catchment area

Methodology used for flow measurement, sampling and calculation

3

Obligatory measurement, sampling and calculation methods were described in the PLC-3 Guidelines. Taking into account that the harmonisation and follow-up of measurement, sampling and calculation is a complicated task, especially for the countries in transition, and that there are still certain unsolved problems, this chapter describes briefly the methods used by the Contracting Parties for better understanding and comparison of the load figures. Detailed information concerning the sampling sites, sampling frequency, parameters and catchment areas is presented in the Annexes to the report.

3.1 Flow

measurement, sampling and calculation of pollution load via rivers

3.1.1 Flow measurement

According to the PLC-3 Guidelines the location of hydrological stations, measurement equipment, frequency of level and flow measurement as well as methods for calculation of annual runoff are regulated by the WMO Guide to Hydrological Practices.

All rivers included in the PLC-3 R of knowledge about the hydro-logical behaviour of the runoff of a comparable neighbouring known river basin.Flow measurements in the territory of each Contracting Party and its degree of conformity with the WMO-Guide are presented below (Table 3.1)

3.1.2 Sampling frequency

According to the PLC-3 Guidelines the sampling regime should be designed on the basis of historical records and should cover the whole flow cycle. The minimum sampling frequency is 12 times per year, appropriately reflecting the expected river flow pattern. The sampling points should correspond to ISO Standards 5667-6 and 5667-9.

Actually the sampling frequency in Denmark, Finland and Germany for large rivers especially concerning organic matter and nutrients, is higher than 12 times per year. The sampling frequency in Poland for all monitored rivers concerning BOD, COD, nutrients and heavy metals is 48-52 times per year (once per week). In the territory of the other Contracting Parties the sampling frequency, as a rule, is 12 times per year.

For partially monitored rivers the frequency is smaller, between 4 and 8 times per year.

The sampling frequency for heavy metals is between 4 and 12 times per year, except for Denmark where water samples are not analysed for heavy metals, AOX etc. In Danish rivers a screening was performed in 1990 with the result that the concentrations for these parameters were below the detection limit. The same problem, i.e. that most of the concentration figures are below the detection limit, occurred in partially monitored rivers in the territories of some other Contracting Parties. Summarised results about sampling frequency are presented in Table 3.2.

3.1.3 Methods for calculation of the riverine load

The methods to be used for calculating load in monitored, partially monitored and unmonitored rivers were described in the PLC-3 Guidelines. According to the Guidelines, selection of the calculation methods would depend on the Contracting Party. The following calculation methods compiled in Table 3.3 were used by the various Contracting Parties.

3.2 Flow measurement, sampling and calculation of pollution load from point sources

3.2.1 Flow measurement

According to the PLC-3 Guidelines, a relative error margin of less than 5% should be the target for open and closed measurement systems in each case. Flow measurement systems and methods should correspond to ISO and DIN standards. Continuous measurement and registration systems should preferably be used. A summary of information about flow measurement from point sources presented by the Contracting Parties has been compiled in Table $3.\overline{4}$.

3.2.2 Sampling methods and sampling frequency

According to the PLC-3 Guidelines, samples from treated and untreated wastewater should always be taken as composite samples. Flowweighted composite samples should be the target. Grab samples are acceptable only in exceptional cases. Sampling frequency depends very much on the polluters. For big polluters sampling frequency is 2-7 times per week. For smaller polluters it is 1-4 times per month, or only a few times per year for very small polluters.

The sampling methods and sampling frequencies presented by the Contracting Parties vary widely. Several Contracting Parties, for instance Germany, used only samples taken by the authorities for pollution load calculations and for this reason the sampling

frequency is only 12 times per year. Other Contracting Parties, for instance Sweden, used all self-control samples as well as samples taken by the authorities for load calculation and for this reason the number of samples is significantly larger. An overview of sampling methods and frequencies is presented in Table 3.5.

3.2.3 Calculation methods

The main calculation methods were presented in the PLC-3 Guidelines. According to the Guidelines, calculated load figures must also include overflows and bypasses. For untreated and unmonit-ored discharge the load may be derived on the basis of per capita load figures. Based on the information provided by the Contracting Parties, an overview of calculation methods is presented in Table 3.6.

3.2.4 Aquaculture

Fish farming plants exist only in Estonia, Finland and Sweden. The load from these plants is calculated based on the amount of nutrients in fish and on the nutrient content of the feed, using for calculation the equations described in the PLC-3 Guidelines. Information about these inputs is included in the source category of industrial plants.

3.2.5 Diffuse inputs from coastal zones into the Baltic Sea

Diffuse inputs from coastal zones into the Baltic Sea include pollutants washed directly into the sea from agricultural areas, managed forests or non-managed natural areas including forests. Information about these inputs is included in the source category of unmonitored rivers.

Table 3.4 Flow measurement for point sources which are reported separately

Table 3.6 Pollution load calculation methods for point sources

* no untreated wastewater

4

Analytical methods and quality assurance

4.1 Analytical methods

4.1.1 Determinants and

analytical methods

The Guidelines for PLC-3 presented descriptions for analytical methods to be used for different determinants in the monitoring programme. In many cases it was recommended that one method be used for the analysis of river water and another for the analysis of wastewater. On the basis of the information given by different countries the analytical methods appear to have corresponded rather well to the Guidelines. In general, the analytical methods used in Denmark, Finland, Sweden and Germany corresponded to Nordic or ISO standards and complied rather well with the requirements in the PLC-3 Guidelines. There was more variation in the methods used by the other countries. For some determinants methods also varied within countries (Annex 1).

The recommended obligatory and voluntary determinants are presented in Table 1.1 (Chapter 1).

In Denmark, Lithuania, Poland and Russia BOD₋ was measured whereas PLC-3 requires BOD_7 to be measured. The results of $BOD₅$ were converted to BOD_r using a factor of 1.15 -1.2.

There was also some variation in methods used to determine nutrient content. In some countries the Nessler method was still used for determination of ammonia and a salicylate method for determination of nitrate (Annex 1). In the determination of total nitrogen and phosphorus the digestion procedure varied. In Poland and in some Danish laboratories the total nitrogen from wastewater was calculated as a sum of nitrogen fractions.

A variety of filters was used to analyse the suspended solids (Annex 1). For this reason, the results for wastewater particularly are not easily comparable with each other, because the amount of suspended solids filtered depends, in general, on the type of filter used as well as on its pore size.

Laboratories have mainly used amalgamation for analysis of mercury and the cold vapour technique for measuring. The enrichment method for mercury by amalgamation was used in Finland, Germany and Sweden, especially in those cases where low mercury values were detected in rivers.

Measurement of metals was generally carried out using atomic absorption spectrophotometry. Whether flame or flameless methods were used depended on the concentration of determinants. For determinations of low metal concentrations ICP/MS was used in Finland, Sweden and Denmark and volta-metry in Germany. The procedures used by the laboratories differed mainly in how well they measured low metal concentrations.

4.1.2 Detection limit

The detection limits depended both on the analytical method used and on the laboratory (e.g. on the quality of the deionized water and reagents and on the equipment used). In most cases it was only possible to approximate the detection limit, because it differed from laboratory to laboratory as well within most countries (Annex 2). Concentrations of determinants are, in general, lower in the rivers of the Nordic countries such as in Finland and Sweden. In these countries it is necessary to use the most sensitive analytical methods. Detection limits varied most in the determination of heavy metals and nutrients.

4.2 Quality assurance

4.2.1 Quality assurance activities

In order to obtain comparable results and to secure reliable data, a quality assurance programme was established before starting PLC-3. In total over 300 laboratories participated in the PLC-3 programme. Because the results from several laboratories were to be combined (Table 4.1), comparability of the results was essential for PLC-3.

The first step was to establish national reference laboratories (NRLs). These were set up in most countries (except Russia) before PLC-3 began. Each country has established two reference laboratories, one for analysis of river water and another for analysis of wastewater.

In the Guidelines it was recommended that the national reference laboratories in all countries take the following steps in order to obtain reliable data for PLC-3:

- set up and carry out internal quality control,
- test new methods when necessary,
- train the personnel of participating laboratories,
- conduct inter-laboratory comparisons.

Various activities of the reference laboratories have been underway both before and during the PLC-3 programme.

4.2.1.1 Internal quality control

The results of inter-laboratory comparisons have shown that the rigour of analytical quality control procedures within a laboratory can be related to its performance. Laboratories should conduct regular controls on the measurements in each batch in order to maintain an acceptable level of accuracy and precision in the analysis. On the basis of the information collected from the national reference laboratories internal quality control appears to have been implemented rather successfully. However it is evident that small laboratories in particular have not always had the resources to introduce quality assurance procedures in accordance with the requirements in the PLC-3 Guidelines.

Quality control procedures require some additional analytical work. Furthermore, unavailability of necessary standards or reference materials have caused problems in implementing quality control procedures. All Swedish laboratories that participated in PLC-3 were accredited.

4.2.1.2 Training

In 1994-1995 the personnel of the national reference laboratories in Estonia, Latvia, Lithuania and Poland participated in international workshops on quality assurance and analytical procedures, e.g. the Quality Assurance Workshop/HELCOM in 1994, the Quality Assurance Workshop/EU-PHARE programme in 1995, the Quality Assurance Workshop/EU-EQUATE programme in 1995 or the workshop on Laboratory Accredition in Sweden in 1995.

When many laboratories in a country have participated in PLC-3, an essential function of the reference laboratory has also been to provide training in analytical procedures and quality assurance. The reference laboratories in Denmark, Estonia, Finland, Latvia, Lithuania and Sweden have provided training for their regional laboratories or for their environmental laboratories participating in PLC-3.

4.2.1.3 International interlaboratory comparisons

The national reference laboratories took part in the international comparisons before or during the PLC-3 programme. The Finnish Environment Institute conducted the comparison test before PLC-3 in June 1994. The samples used were artificial and surface water, municipal wastewater or wastewater from the pulp and paper, and metal industries. The main obligatory determinants were compared. The results of the comparisons are summarized in Table 4.2.

The results showed that the relative standard deviation (between laboratories) was fairly high in some cases; e.g. up to 30-45% in the analysis of nitrogen compounds, several heavy

metals (Cr, Cu, Pb, Ni, Zn) and mercury. Generally deviation was greater in determination of low concentrations. In this comparison the target value for criterion of bias was 10-40% and it also varied according to the concentration and on the determinant. The results were regarded as acceptable, if they differed less than $10-40\%$ from the mean value or the assigned value (Table 4.2).

The following reasons seem to be the most significant sources of errors:

- 1. lack of appropriate resources (facilities, equipment, labware, pure ionized water, pure chemical reagents),
- 2. lack of systematic quality control procedures,
- 3. the use of methods which may not provide good quality data.

The comparisons were conducted only shortly before PLC-3 was started. In general reference labor-

atories had neither the time nor the financial resources to effect many improvements. While PLC-3 was underway some improvements were made, particularly in the reference laboratories of Estonia, Latvia and Lithuania with the support of the EU-PHARE programme. How-ever, in 1995 the laboratories received mainly educational rather than technical support.

Some reference laboratories of Estonia, Latvia, Lithuania and Poland also participated in the comparisons conducted within the EU/EQUATE project in June 1995, in which most of the obligatory determinants in PLC-3 were compared. In general the results of this comparison were acceptable.

The Russian laboratories participating in PLC-3 did not take part in the international comparisons. Therefore it is impossible to compare Russian results with those from other countries.

4.2.1.4 National interlaboratory comparisons

In the PLC-3 Guidelines the national reference laboratories were asked to conduct national inter-laboratory comparisons to monitor the performance of other laboratories in the country.

The countries conducted the inter-laboratory comparisons before or during PLC-3 (Table 4.3). There were some flaws in the way of the interlaboratory comparisons. In some cases only a few determinants were compared. In other cases, only a few sam-

ples were distributed; the concentrations of the types of samples compared did not correspond to the real samples. In some countries these comparisons were also constrained by low financial resources.

National inter-laboratory comparisons were not necessary in Germany because only two national reference laboratories participated in PLC-3.

In terms of the results of BOD $_{\gamma'}$ the outcome in different countries seemed to be quite similar, whereas there were large differences between the countries when COD_{c} values were compared (Table 4.4). AOX and TOC, which were compared by Sweden, Finland and Denmark, seemed to be quite reliable determinants.

Concentrations of nitrogen compounds in the samples distributed by the different countries varied significantly. The variation in results generally depended on the concentration of the samples. In this data the results varied between 5-35% and the variation was the highest in the determination of ammonium nitrogen and total nitrogen. In general, the results of total phosphorus varied between 5-20%.

Determination of suspended solids was only compared by Denmark, Poland and Sweden. The Swedish laboratories used different kind of filters for determination of suspended solids in surface water. The observed variation in the Swedish comparison test, 35-45%, corresponded quite significantly with the variation in results reported by the countries in PLC-3. For higher values, variation was much lower.

Determination of mercury was compared only by Finland and Sweden. In this data the results varied up to 30% $(Hg < 0.2 \,\mu g/l)$. In determination of low

concentrations of Hg from surface waters the variation can be much higher, up to 60%.

Heavy metal concentrations in the distributed samples varied considerably. The results varied up to 65%. The variation was the highest in determination of heavy metal concentrations at a level ≤ 1 μ g/l in surface waters.

The variation in results of the national comparisons did not differ very much from the variation observed in the international comparisons conducted for the national reference laboratories before the start of PLC-3 (Table 4.4). The variation between reference laboratories was 5-10% lower than the variation obtained in the national comparisons for determination of BOD₇, COD_{cr} and heavy metals.

4.3 Conclusions

After the PLC-2 report has been published it became clear that a quality assurance system had to be established before PLC-3 got underway. Within the framework of such a system, laboratories could improve their ability to provide data of appropriate quality. The national reference laboratories have played an essential role in improving analytical performance in most countries. The laboratories have provided personnel training for their regional or industrial laboratories and wastewater treatment plants. In addition, most laboratories participating in PLC-3 carried out internal quality control during the programme, but some had difficulties in carrying out the national comparisons. Consequently some problems have

arisen in drawing conclusions about the performance of the laboratories in some countries. Unfortunately some countries have had problems related to the use of insensitive and/or inappropriate methodology, lack of good quality labware and reagents and lack of adequate instrumentation.

In general the national reference laboratories took part in the international comparisons. However, there have been some problems in controlling the quality of the work that has been carried out by their national laboratories. Overall the laboratories have worked towards improvement of the quality of the data in PLC-3 and they have been aware of the importance of quality assurance procedures. Through better quality control the measurements are becoming more reliable. In addition, many laboratories participating in PLC-3 have implemented the quality system based on the EN 45001 and the ISO/IEC Guide 25. In any case, the quality assurance programme has to some degree supported the data produced for PLC-3, although it takes time to achieve improvements in quality assurance and analysis, especially in countries where financial resources are a problem. Nevertheless, the programme has provided essential information on methodology and quality assurance for many laboratories.

In order to obtain relevant and reliable data in the next stages of PLC, it is essential that the whole analytical process should proceed under a well-established quality assurance programme in PLC laboratories.

*Only one sample was distributed

** A high concentration of determinant

5

Pollution load entering the Baltic Sea in 1995

5.1 General information about rivers, municipalities and industrial plants

PLC-3 examined nutrient, organic matter and heavy metal loads entering the Baltic Sea marine environment from rivers, coastal areas and direct point sources such as municipalities and industrial plants. However, point source discharge was only taken into account when the source was discharging directly into the Baltic Sea or when it was located downstream of the hydrological/ hydrochemical station in the river for which the load is given. Therefore, only a small portion of the total point sources located within a river basin were considered, which means that an inventory of all point sources in the whole Baltic Sea catchment area is lacking.

The runoff and direct point source discharge considered in PLC-3 is 466 320 million m ∇/a , of which about 93% represents runoff into the Baltic Sea from monitored rivers. The runoff from coastal areas including unmonitored rivers is the second biggest share with nearly 6%. About 1% of the discharge came from municipalities and industrial plants, to which treated and untreated municipal discharge and treated industrial discharge each contributed about 0.3%. The amount of untreated direct industrial discharge was negligible in comparison with all other pollution sources (see Figure 5.1). In Figures 5.2 and 5.3 the distribution of runoff and amount of direct point source discharge from all pollution source categories considered in PLC-3 is shown by subregion and by Contracting Party, respectively.

Sea in 1995 by Contracting Party

5.1.1 Information about

rivers

Riverine load consists of load from different pollution sources within a river catchment area such as industrial plants, municipal wastewater treatment plants, diffuse load (e.g. agriculture, forestry, and scattered dwellings) and natural background load. It should also be noted that the load for transboundary rivers contains not only pollution load from the country in which the measurement site is located, but also load from point and diffuse sources situated in the territory of other Contracting Parties and/or Non-Contracting Parties.

In 1995 263 rivers were monitored or partially monitored. The total runoff of these rivers amounted to 434 690 million m ∇/a with a total catchment area of 1.5 million km , of which 1.2 million km \Box are located within the territory of Contracting Parties in which the monitoring stations are situated. The runoff from coastal zones and unmonitored rivers amounted to 26 190 million m \sqrt{a} (6%) with a corresponding catchment area of nearly 125000 km . According to Figure 5.4 and Table 5.1 riverine runoff into the Gulf of Finland was the greatest, with nearly 100 000 million m ∇/a from a total catchment area of monitored rivers of nearly 400 000 km . The runoff into the Bothnian Bay, the Bothnian Sea and the Baltic Proper amounted in each case to about 90 000 million m ∇/a . The total monitored river catchment area of the Baltic Proper subregion was the largest, nearly 500 000 km ^{\right}. The total monitored river catchment area of the Bothnian Bay and the Bothnian Sea subregions amounted to approximately $200\,000$ km \Box . The number of monitored rivers varied: There were 16 in the Bothnian Sea catchment area, 21 in the Bothnian Bay catchment area, 39 in the Gulf of Finland catchment area and 49 in the Baltic Proper catchment area. The seven largest rivers flowing into the Baltic Sea, all but the Göta älv were situated in these four subregion catchment areas. Most of the monitored rivers were located in the Western Baltic catchment area, but the runoff from these 73 rivers was very limited. These rivers are streams compared with the large rivers in the catchment area of the other subregions. The Sound had the

lowest runoff, about 220 million m ∇/a , and the smallest catchment area, nearly 1000 km .

According to Figure 5.5 and Table 5.2, the largest portion of the runoff - 174 330 million m \sqrt{a} , which is nearly three times higher than the runoff from Russia (69 320 million m[]/a), Finland (63 110 million m \sqrt{a}) and Poland (54 670 million $m\sqrt{a}$ - entered the Baltic Sea from the Swedish Baltic Sea catchment area. Sweden has monitored 42 rivers, Russia 19 rivers, Finland 26 rivers and Poland 12 rivers. Sweden is also the country with the largest Baltic Sea river catchment area, about 420 000 km₁. River catchment area in Poland, Russia and Finland amounts to slightly over 300 000 km_| in each case. Denmark and Germany have the smallest river catchment area, about 15 000 km \Box and also the lowest amount of riverine runoff. In Denmark 103 rivers were monitored, with a runoff of 4 340 million m ∇/a . Germany monitored 36 rivers with a runoff of 3 690 million m ∇a .

5.1.2 Information about municipalities

It should be stated, that the 177 million m /a wastewater originating in the Kaliningrad region of Russia with a population of 900 000 and flowing into the Baltic Proper, of which 151 million m ∇ a were untreated, could not be taken into account in the report because no information about municipalities and loads was submitted by Russia. It is also important to stress, that due to the fact that Russia only submitted the total amount of direct municipal wastewater discharges into the Gulf of Finland, it was impossible to present the distribution of Russian municipalities in Figures 5.11 and 5.12.

In 1995 the total amount of direct municipal wastewater discharge was 3 490 million m ∇/a , from 15 million inhabitants connected to a total of 430 municipalities. Approximately 13 million inhabitants were connected to 155 municipal wastewater treatment plants $(MWWTPs) > 10000$ PE (Population Equivalents) which produced nearly 2 950 million m ∇/a treated wastewater. The treated wastewater discharge from 264 small settlements (including non-

systematically monitored settlements from Poland) with a total of 377 000 inhabitants was only 55 million m ∇a , which is less than 2% of total direct municipal discharge. Nearly 500 million $m\sqrt{a} (14\%)$ of the municipal wastewater discharges into the Baltic Sea were untreated. It was produced by 10 municipalities with a total of 1.6 million inhabitants (see Figure 5.6). According to Figures 5.7 and 5.8 the largest share of the untreated municipal wastewater, 430 million m_{Va}, originated in Saint Petersburg and in the Leningrad region of Russia with a population of 1.2 million, flowing into the Gulf of Finland. The remaining 55 million m ∇/a of direct untreated municipal wastewater discharges came from Poland (39 million m[]/a), Latvia (16 million m[]/a), Lithuania (0.4 million m \sqrt{a}) and Estonia (0.2 million m/d) and flowed into the Baltic Proper, the Gulf of Riga and the Gulf of Finland. None of the other Contracting Parties discharged untreated municipal wastewater into the Baltic Sea; thus no such discharge could enter the Baltic Sea from the catchment areas of the Bothnian Bay, the Bothnian Sea, the Archipelago Sea, the Western Baltic, the Sound or the Kattegat.

The distribution of direct municipal wastewater discharges, population and number of municipalities by subregion and by Contracting Party is shown in Figures 5.7, 5.9 and 5.11 and in Figures 5.8, 5.10 and 5.12, respectively. The largest share of treated municipal wastewater came from 14 MWWTPs > 10 000 PE discharging directly into the Gulf of Finland, treating 1 550 million $m\sqrt{a}$ of wastewater and serving 5 million inhabitants. The treated discharge from 44 small settlements with a population of 23 100 discharging directly into the Gulf of Finland amounted to 5 million $m\sqrt{a}$. The second largest share of treated municipal wastewater, 500 million m \Box/a , is produced by 48 MWWTPs > 10 000 PE discharging directly into the Baltic Proper and serving 3.8 million inhabitants. The treated direct discharge into the Baltic Proper from 85 small settlements amounted to 19 million m ∇/a with a total population of 155 000. The direct treated municipal wastewater discharge from MWWTPs > 10 000 PE discharging directly into the Sound and the Kattegat amounted in each case to

Table 5.1 Runoff from rivers and coastal areas including corresponding catchment areas entering the Baltic Sea from the subregion's catchment area in 1995 (see Abbreviation List)

Table 5.2 Runoff from rivers and coastal areas including corresponding catchment areas entering the Baltic Sea from the Contracting Party's catchment area in 1995

 $N.I. = No information$

 $=$ This source does not exist.

about 230 million m ∇/a , serving a total of 1.4 million inhabitants. The lowest municipal discharges from MWWTPs > 10 000 PE originated in Latvia and Lithuania and were discharged into the Archipelago Sea, the Bothnian Bay and the Bothnian Sea.

In Denmark, Germany, Finland and Sweden all municipal effluents were treated in municipal wastewater treatment plants. Nearly all of these plants used mechanical, chemical and biological treatment methods with phosphorus removal rates of between 80% and 97%. In all the Danish plants except that of Copenhagen (LYN), nitrogen removal also took place, with elimination rates between 70% and 99%. In 10 of the 24 German plants additional nitrogen removal took place with elimination rates of about 80% to 98% . In the Finnish and Swedish plants the nitrogen removal rate was generally less than 50%, except in 4 of the 55 Swedish plants where the nitrogen removal rate was of the order of 70%. In one of the Polish plants and 2 of the Estonian plants the phosphorus removal rate was over 80%.

5.1.3 Information about industrial plants

The 183 industrial plants considered in PLC-3 discharged 1 950 million m[/a wastewater directly into the Baltic Sea in 1995, of which more than 1948 million m_//a was treated wastewater discharged by 177 industrial plants. The untreated wastewater discharge from 6 industrial plants amounted to 0.42 million m ∇ a, which is less than 0.1% of total direct industrial discharge. One of these industrial plants discharging untreated wastewater directly into the Baltic Sea is located in Estonia, two are located in Latvia and three in Poland.

The distribution of direct industrial wastewater discharges and number of industrial plants by subregion and by Contracting Party is shown in Figures 5.13 and 5.15, and Figures 5.14 and 5.16, respectively. According to these Figures the largest share of treated industrial wastewater was discharged by 60 industrial plants into the Bothnian Bay (540 million m \sqrt{a}), the Kattegat $(460$ million m ∇/a) and the Bothnian Sea (350 million m ∇/a). 13 industrial plants

Total riverine runoff: 460 880 million m $^3\!/\!a$ Figure 5.4 Riverine runoff into the Baltic Sea in 1995 by subregion

Total riverine runoff: $460\,880$ million m 3 /a Figure 5.5 Riverine runoff into the Baltic Sea in 1995 by Contracting Party

 \blacksquare untreated direct discharge into the Baltic Sea from municipalities and small settlements

Intreated direct discharge into the Baltic Sea from small settlements

I treated direct discharge into the Baltic Sea from MWWTP>10 000 PE

Figure 5.8 Distribution of direct municipal discharge by Contracting Party into the Baltic Sea in 1995

 \Box untreated direct discharge into the Baltic Sea from municipalities and small settlements I treated direct discharge into the Baltic Sea from small settlements

Internet direct discharge into the Baltic Sea from MWWTP>10 000 PE

 $\hfill \textbf{L}$ inhabitants connected to municipalities and small settlements without wastewater treatment discharging directly into the Baltic Sea
Inhabitants connected to small settlements with treatment plants discharging

directly into the Baltic Sea

Trinhabitants connected to MWWTP > 10 000 PE discharging directly into the **Baltic Sea**

Figure 5.9 Distribution of population by subregion connected to municipalities discharging directly into the Baltic Sea in 1995

discharged 280 million m ∇/a treated wastewater directly into the Gulf of Finland and 56 industrial plants discharged 250 million m ∇/a treated wastewater directly into the Baltic Proper. Treated wastewater from 48 industrial plants discharging directly into the Sound, the Western Baltic and the Gulf of Riga were approximately 40 million m ∇a .

The distribution of the amount of treated industrial wastewater produced by the nine branches of industry considered in PLC-3 is given in Figure 5.17 by subregion and in Figure 5.18 by Contracting Party. The largest share of the direct industrial wastewater discharge, 890 million m ∇/a , came from 40 pulp and paper plants located in Finland, Russia, Latvia, Poland and Sweden. The main wastewater discharge emitted by this branch of industry was introduced from the 24 Swedish (420 million m ∇ a) and the 9 Finnish (320 million m /a) pulp and paper plants. The second largest amount of industrial wastewater, 530 million m \Box/a , was discharged by "other industry", of which 425 million $m\sqrt{a}$ came from 25 Danish plants. The remaining 100 million m γ a were emitted by industrial plants of this branch of industry in Latvia, Lithuania, Poland and Russia. Wastewater discharges from the chemical industry amounted to 230 million m ∇ a with the largest share, 130 million m ∇/a , from Poland and 90 million m ^{α} from Finland. Wastewater from the iron and steel industry of 195 million m /a was discharged into the Bothnian Bay and into the Baltic Proper. One of these plants, discharging 140 million m ∇a , is situated in Finland, two of them, discharging 54 million m \sqrt{a} , are located in Sweden and the remaining 99 million m_{\sqrt{a}} came from Poland. One Swedish metal enrichment plant discharged 27 million m ∇ a wastewater into the Bothnian Bay and two Finnish non-ferrous metal plants discharged 21 million m ∇a wastewater into the Bothnian Bay. There are also 8 petrochemical plants, discharging 34 million m ∇/a , of which 21 million m ∇/a were discharged by the two Finnish plants. All Contracting Parties except Finland and Lithuania have food industry discharging directly into the Baltic Sea, consisted of a total of 39 plants producing 23 million m ∇/a wastewater. The largest share of this wastewater

Total amount of treated wastewater from 177 industrial plants: 1 948 190 \cdot 10 $^{\rm 3}$ m $^{\rm 3}/$ a Figure 5.14 Distribution of direct industrial discharge by Contracting Party into the Baltic Sea in 1995

Figure 5.15 Distribution of industrial plants by subregion discharging directly into the Baltic Sea in 1995

47

from the food industry, 18 million m ∇a , were discharged from 27 Danish plants. The leather and textile industry is the branch producing the smallest amount of wastewater, 0.3 million m ∇/a , which was discharged from one Russian plant into the Gulf of Finland.

5.2 Organic matter load going into the Baltic Sea

Organic matter is one of the concerns of the marine environment. Monitoring of oxygen depletion in waterbodies started already in the 1920s. Lack of oxygen is in some areas a problem for the open sea, especially in bottom layers in deep parts of the sea as well as in some coastal zones.

According to the PLC-3-Guidelines the organic matter load is measured as $\mathrm{BOD}_{7'}\mathrm{COD}_{\mathrm{Mn'}}\mathrm{COD}_{\mathrm{Cr}}$ or as TOC. In the following, however, only the results for BOD₇ are given*,* due to the fact that this parameter was measured in nearly all Contracting Parties for most pollution sources. In Finnish and Swedish rivers the BOD₇ load was calculated on the basis of TOC. In Denmark the BOD₇ load was calculated on the basis of the BOD_5 . By that, it is possible to give an overview about the BOD₇ load by subregion and by Contracting Party.

In 1995 the total BOD₇ load going into the Baltic Sea amounted to 1 140 080 t. The distribution of riverine and direct point source BOD_7 load is given in Figure 5.19. The major part of organic matter load, 80%, entered the Baltic Sea via rivers and coastal areas, of which 76% was via monitored rivers and 5% via unmonitored rivers and coastal areas. The BOD $_{_{\gamma}}$ load from municipalities and industrial plants discharging treated wastewater directly into the Baltic Sea is 9% in each case. The share of BOD_7 load from these pollution sources is considerably higher than the share of the amount of wastewater (approximately 0.3% in each case; see chapter 5.1 and Figure 5.1). The share of untreated municipal BOD₇ load was quite low, only 1%, but it should be noted that the untreated portion of the load from

the Russian Kaliningrad region discharging directly into the Baltic Proper is missing. The share of untreated direct industrial BOD₇ load was also quite low, 0.002%. However, the actual industrial load should be higher due to the fact that in some countries (for instance Russia and Estonia) the main industries are connected to municipal wastewater treatment systems and were therefore not reported separately.

In Tables 5.3 and 5.4 the distribution of BOD₂ load for the 6 pollution source categories is given by subregion and by Contracting Party, respectively. Mostly the rivers are the dominant source of BOD $_{_{\gamma}}$ load. However, municipal and industrial BOD_7 load discharging directly into the Bothnian Sea, the Western Baltic and the Kattegat amounted to 30% in each case. In Denmark this share of $\mathrm{BOD}_7 \mathrm{load}$ increased to 50% , but it should be noted that Danish figures for unmonitored rivers also included load from municipalities and industrial plants.

In 1995 up to 43% of the total BOD7 load was discharged into the Baltic Sea from the Baltic Proper catchment area. The main part of the $BOD₇$ load entered the Baltic Proper via the three largest rivers: the Vistula (164 620 t/a), the Oder (87 640 t/a) and the Nemunas (91 880 t/a). The $BOD₇$ load from these three rivers, running through the most densely populated parts of the Baltic Sea

catchment area, comprised approximately 30% of the total BOD₇ load, but the corresponding runoff was only about 15% of the total riverine runoff. The second largest share of BOD_7 load, 237 350 t/a, entered the Baltic Sea from the Gulf of Finland catchment area, of which 125 900 t/a were discharged by the Neva and 45 000 t/a by direct municipal and industrial discharges from Saint Petersburg and the Leningrad region, which still contains some untreated wastewater.

The organic matter load from the rivers discharging into the northern part of the Baltic Sea (Bothnian Bay and Bothnian Sea) is mostly from natural areas with low impact from human activity caused by the high content of humic matter in these river waters (from forest, peat-soils etc.). In fact, the rivers discharging into the Bothnian Bay have $BOD₇$ concentrations below the detection limit, so that the $BOD₇$ load figures were calculated on the basis of measured TOC values.

The municipal $\mathsf{BOD}_7^{}$ load into the Bothnian Bay, the Bothnian Sea, the Archipelago Sea, the Western Baltic, the Sound and the Kattegat is low due to effective treatment of municipal wastewater in Finland, Sweden, Germany and Denmark, where the BOD $_{_7}$ removal rate, as a rule, is higher than 90%. In

load going into the Baltic Sea in 1995 by subregion

 $-$ = nothing to report (this source does not exist)

l) Riverine BOD₇ load in Finland and Sweden has been calculated

 * $=$ data not available (should have been reported) $# =$ data not available, but not obligatory

 from the TOC load using the coefficient 0.145. 2) Danish BOD $_{_{7}}$ has been calculated from the BOD $_{_{5}}$

data not complete

using the coefficient 1.15.

 $-$ = nothing to report (this source does not exist)

l) Riverine BOD₇ load in Finland and Sweden has been calculated

 $* =$ data not available (should have been reported) $# =$ data not available, but not obligatory data not complete

 from the TOC load using the coefficient 0.145. 2) Danish BOD₇ is has been calculated from the BOD₅ using the coefficient 1.15.

Figure 5.20 BOD₇ load going into the Baltic Sea in 1995 by subregion

Figure 5.21 BOD₇ load going into the Baltic Sea in 1995 by Contracting Party

addition to Saint Petersburg and the Leningrad region high amounts of municipal BOD, load entered the Baltic Sea from the large cities in countries in transition on the coast of the Baltic Proper subregion, where the efficiency of wastewater treatment is not optimal.

Only in the case of the Bothnian Sea the industrial organic matter load is higher than the municipal organic matter load. The main industries discharging into that area are pulp and paper mills. The highest amount of organic matter load (BOD and COD) into the Bothnian Bay, the Bothnian Sea and the Gulf of Finland by industry also originated from the pulp and paper branch, mainly from Sweden and Finland. In Finland all the plants have biologically activated sludge removal treatment systems, but in Sweden some plants still rely on mechanical wastewater treatment methods.

5.3 Nutrient load going into the Baltic Sea

The main problem of the Baltic Sea Area is the nutrient load. Since the turn of the century, the Baltic Sea has changed from an oligotrophic clear-water sea into a highly eutrophic one (LARSON, 1985). The state of the Baltic Sea is alarming in many of its subregions: these regions have become overloaded with nutrients. Nitrogen and phosphorus as such do not pose any direct hazards to marine organisms or people. Excessive nutrient inputs may disturb the balance of the ecosystem. Excessive primary production, caused by high concentrations of phosphorus and nitrogen, has caused algae blooms, especially of the bluegreen variety, to proliferate in the Baltic Sea. The abundance of toxic algae populations has even increased.

5.3.1 Nitrogen load going into the Baltic Sea

In 1995 the total waterborne N_{total} input into the Baltic Sea amounted to

760 750 t. The distribution of this load among the different pollution sourcecategories is given in Figure 5.22. The major part of N_{total} load, 90%, entered the Baltic Sea via rivers, 76% coming via monitored rivers and 14% via unmonitored rivers and coastal areas. The treated municipal and industrial share of

 N_{total} load discharging directly into the Baltic Sea comprised 8% and 2%, respectively. The portion of untreated municipal and industrial N_{total} load discharging directly into the Baltic Sea is quite low, only 0.3% and 0.2%, respectively. The calculations in PLC-3 can be considered more reliable and precise than calculations in PLC-1 and PLC-2, but still many uncertainties remain due to incomplete data sets especially from the Russian Baltic Sea catchment area, e.g. from the Kaliningrad region (see Tables 5.5 and 5.6).

In Tables 5.5 and 5.6 the N_{total} load distribution among the 6 pollution source categories is given by subregion and by Contracting Party. The part of direct municipal and industrial nitrogen inputs into all the Baltic Sea subregions is quite low: less than 20%. Into most of the subregions' direct municipal nitrogen inputs are of much greater importance than the corresponding industrial inputs. The only subregion, to which industrial nitrogen inputs exceeded the municipal nitrogen inputs is the Bothnian Bay, where the industrial nitrogen load originated primarily from the pulp and paper industry in the coastal areas of Finland and Sweden.

In 1995 up to 42% of the N_{total} load was discharged into the Baltic Sea from the Baltic Proper catchment area. The major part of this load entered the Baltic Proper via the three largest rivers: the Vistula (112 800 t/a), the Oder (76 970 t/a) and the Nemunas (34 190 t/a). The nitrogen load from these three rivers comprised approximately 30% of the total N_{total} load, but the corresponding runoff was only about 15% of the total riverine runoff. The second largest share of Ntotal load, 132 900 t/a (17%), entered the Baltic Sea from the Gulf of Finland catchment area, of which 54 170 t/a and 24 950 t/a being discharged by the Neva and the Narva, respectively. The riverine nitrogen inputs into all other subregions are lower. The Gulf of Finland (25 390 t/a) and the Baltic Proper subregions (15 600 t/a) also received the highest amounts of N_{total} load from urban areas. Especially the municipal discharges from Saint Petersburg and the Leningrad region into the Gulf of Finland constituted the main part of the N_{total} load, 19 680 t/a. 1 784 t/a of the untreated municipal effluents were discharged directly into the Baltic Sea from Poland. The main treated industrial nitrogen discharges are going into the Gulf of Finland (26%), the Baltic Proper (18%) and the Bothnian Sea subregions (17%). Nearly all the untreated industrial effluents, 1 570 t/a, were discharged from Estonian industry directly into the Gulf of Finland.

Table 5.5 Riverine and direct point source N_{total} load going into the Baltic Sea in 1995 by subregion

 $-$ = nothing to report (this source does not exist)

 $\mathrm{*} =$ data not available (should have been reported)

 $# =$ data not available, but not obligatory

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Table 5.6 Riverine and direct point source Nitrogen load going into the Baltic Sea in 1995 by Contracting Party

 $-$ = nothing to report (this source does not exist)

 $* =$ data not available (should have been reported) $# =$ data not available, but not obligatory

Figure 5.23 N_{total} load going into the Baltic Sea in 1995 by subregion

5.3.2 Phosphorus load going into the Baltic Sea

In 1995 the total waterborne P_{total} load going into the Baltic Sea amounted to 37 650 t. The distribution of this load among the different pollution source categories is given in Figure 5.25. The major part of $\mathrm{P_{total}}$ load, 81%, entered the Baltic Sea via rivers, 72% via monitored rivers and 9% via unmonitored rivers and coastal areas. The treated municipal and industrial share of P_{total} load discharging directly into the Baltic Sea comprised 13% and 5%, respectively. The portion of the untreated municipal and industrial P^{total} load discharging directly into the Baltic Sea is quite low, only 0.1% and 0.02%, respectively. The calculations in PLC-3 can be considered more reliable and precise than calculations in the two previous load compilations, but still many uncertainties remain due to incomplete data sets especially from the Russian Baltic Sea catchment area, e.g. from the Kaliningrad region (see Tables 5.7 and 5.8).

In Tables 5.7 and 5.8 the P_{total} load is distributed among the 6 pollution source categories and given by subregion and by Contracting Party, respectively. The portion of direct municipal and industrial phosphorus inputs from all the subregions going directly into the Baltic Sea is quite low: less than 20%. Only in Denmark the treated municipal and industrial discharges going directly into the Baltic Sea were equal in quantity to the riverine inputs. In the Sound, however, direct municipal and industrial phosphorus load constituted 72% of the total phosphorus load. In most of the subregions, municipal phosphorus inputs discharging directly into the Baltic Sea are of much greater importance than the corresponding industrial inputs. However, into the Bothnian Bay, the Bothnian Sea and the Archipelago Sea the direct industrial phosphorus inputs were more than triple the direct municipal phosphorus inputs. These industrial phosphorus inputs came primarily from the pulp and paper industry in the coastal areas of Finland and Sweden. In addition, in the Archipelago Sea, the intensive fish farming contributed 18% of the total industrial phosphorus load into that subregion.

In 1995 up to 47% of the total P_{total} load entered the Baltic Sea from the Baltic Proper catchment area. The main part of this load was discharged into the Baltic Sea via the three largest rivers, the Vistula (7 320 t/a), the Oder (4 920 t/a) and the Nemunas (1 230 t/a). These three rivers together contributed approximately 36% of the $\rm P_{total}$ load, while total runoff from these three rivers constituted about 15% of the total riverine runoff. The second largest share of P_{total} load entered the Baltic Sea from the Gulf of Finland catchment area, 22% or 8 160 t/a, of which 2 800 t/a and 670 t/a were discharged by the Neva and the Narva, respectively. The riverine phosphorus load into the Archipelago Sea, the Western Baltic, the Sound and the Kattegat was considerably lower. The Gulf of Finland (2 610 t/a) also received the highest

amount of P_{total} load from urban areas, of which direct municipal discharge from Saint Petersburg and the Leningrad region made up the largest share, 2 440 t/a. 74% of all untreated municipal effluents (290 t/a) was discharged directly into the Baltic Sea by Polish municipalities. Most of the treated industrial phosphorus load was discharged into the Gulf of Finland (41%), the Baltic Proper (21%) and the Bothnian Sea subregions (13%), and originated from Russia (45%), Sweden (17%) and Finland (14%). Only Estonia, Latvia and Poland discharged un-treated industrial phosphorus load directly into the Baltic Sea but it was negligible compared to the inputs from all other pollution source categories.

Table 5.7 Riverine and direct point source P \Box load going into the Baltic Sea in 1995 by subregion

 $=$ nothing to report (this source does not exist)

 $\mathrm{*} =$ data not available (should have been reported)

 $# =$ data not available, but not obligatorY

Table 5.8 Riverine and direct point source phosphorus load going into the Baltic Sea in 1995 by Contracting Party

 $- =$ nothing to report (this source does not exist)

 $\text{*} =$ data not available (should have been reported)

 $\# =$ data not available, but not obligatory

Figure 5.26 P_{total} load going into the Baltic Sea in 1995 by subregion

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Figure 5.27 P_{total} load going into the Baltic Sea in 1995 by Contracting Party

5.3.3 Nutrient concentrations in monitored rivers

To present and compare concentrations in the monitored PLC-3 rivers they were divided into three groups according to their flow rate (Q) in 1995. The number of rivers in the Baltic Sea catchment area of each Contracting Party, in each of the three flow rate groups, is shown in Table 5.9:

- a) small rivers: streams with flow rate $≤ 1 m³/s$
- b) medium size rivers: rivers with flow rate between 5 and 50 $\mathrm{m}^{\mathrm{3}}/\mathrm{s}$
- c) large rivers: rivers with flow rate $> 50 \text{ m}^3/\text{s}$.

Rivers belonging to the following 4 subregion groupings of the Baltic Sea were analysed:

Within each flow rate category in each of the four subregion groupings the monitored rivers were further grouped according to their flow-weighted nitrogen and phosphorus concentration, which was calculated by dividing the nitrogen or phosphorus load by the corresponding flow rate, respectively. Three categories of nitrogen concentrations and phosphorus concentrations were defined:

Nitrogen:

concentration lower than 1 mg/l N concentration from 1 to 3 mg/l N concentration higher than 3 mg/l N

Phosphorus: concentration lower than 0.050 mg/l P

Table 5.9:Number of rivers in the Baltic Sea catchment area of each Contracting Party in

each of the three flow rate groups

concentration from 0.050 to 0.150 mg/l P concentration higher than 0.150 mg/l P

The result of grouping the flowweighted nitrogen and phosphorus concentrations is shown in Figure 5.28. Nearly 95% of the rivers with a mean flow \leq 5 m[$\frac{1}{s}$ (e.g. small rivers) discharging into BAP (A) and more than 90% of rivers discharging to WEB+KAT+SOU (D) have flow-weighted nitrogen concentrations higher than 3 mg/l N (and up to 25 mg/l N). On the other hand, all small streams discharging into BOB+BOS+ARC (B) and 50% of the small rivers discharging into GUF+GUR (C) have flow-weighted nitrogen concentrations lower than 3 mg/l N. The large rivers (defined as rivers with mean flow > 50 m[$\frac{1}{s}$) have on average lower flowweighted nitrogen concentrations than the small rivers in each of the four defined subregion groupings of the Baltic Sea. Only in subregion A (BAP) more than 50% of the large rivers have flowweighted nitrogen concentrations higher than 3 mg/l N. For medium-sized rivers (those with mean flow ranging between 5 and 50 m $\sqrt{ }$ s) the distribution of flowweighted nitrogen concentration is quite similar to that of the large rivers in the four subregion groupings of the Baltic Sea, except for subregion grouping D (WEB+KAT+SOU) where the distribution is comparable to that of the small rivers. Further, in the group of mediumsized rivers the widest range of flowweighted nitrogen concentrations appears.

The distribution described for flow-weighted nitrogen concentrations in the three groups of rivers is also valid for the corresponding phosphorus concentrations (Figure 5.29). The group "small rivers" contains the greatest number of rivers with high flow-weighted phosphorus concentrations and the group "large rivers" has the greatest number of rivers with low flow-weight-

ed phosphorus concentrations. Subregion groupings A (BAP) and D (WEB+KAT+SOU) have the highest number of rivers with high flowweighted phosphorus concentrations, whereas more than 95% of the rivers have flow-weighted phosphorus concentrations higher than 0.050 mg/l P with maximum values of 0.600 mg/l P. In subregion groupin B(BOB+BOS+ARC) only 25% of the small rivers have flowweighted phosphorus concentrations higher than 0.050 mg/l P. Only in subregion grouping A (BAP) there are large rivers with flow-weighted concentrations higher than 0.150 mg/l P (30%). In the other subregion groupings more than 60% of the large rivers have flowweighted phosphorus concentrations lower than 0.050 mg/l P. The highest range of flow-weighted phosphorus concentrations occurs in the group of medium-sized rivers.

Large rivers generally contain a higher amount of groundwater leakage than small rivers, leading to lower concentrations in the former group. Nutrient concentration in many large rivers is also reduced by nutrient retention and turnover in lakes and in the rivers, in wet meadows and on periodically flooded riparian areas. Furthermore, these selfpurification processes will reduce nutrient fluxes to the Baltic Sea significantly. These self-purification processes are markedly reduced where the rivers have been regulated by straightening or surrounded by dikes and where the riparian areas have been drained. Natural processes of retention and turnover of nutrients have a minor impact on nutrient load in small rivers. In addition, most of the small rivers and streams are located in Germany, Denmark and southern Sweden in catchment areas where farming is very intensive. Therefore the highest percentage of rivers with high flow-weighted nitrogen and phosphorus concentrations is found among the small rivers draining subregion groupings D (WEB+KAT+SOU) and A (BAP) and the lowest percentage draining subregion grouping B (BOB+BOS+ARC), where the amount of arable land is low in general. Many of the large rivers draining catchment areas situated in the central and northern parts of Sweden and Finland flow through areas with low percentages of arable land.

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5.3.4 Area-specific nutrient load in 1995

Area-specific nutrient load is calculated as the total nutrient load from all pollution sources divided by the corresponding catchment area. The highest areaspecific nutrient load of nitrogen occurs in the catchment area of the Sound (3 153 kg N/km²), the Western Baltic (2168 kg N/km²) and the Kattegat (860 kg N/km²) (Table 5.5). The highest corresponding figures by Contracting Party are for Denmark (2 208 kg N/km²), Germany (1 288 kg N/km²) and Latvia (592 kg N/km²) (Table 5.6). The highest area-specific nutrient load of phosphorus occurs in the catchment area of the Sound (220 kg P/km²), the Western Baltic (63 kg P/km²) and the Archipelago Sea (54 kg P/km²) (Table 5.7). The highest corresponding figures by Contracting Party are for Denmark (83 kg P/km²), Poland (43 kg P/km²) and Germany (35 kg P/km²) (Table 5.8). High areaspecific nitrogen load is related to the high rate of agricultural activity, such as large amounts of live-stock per unit area and the use of large quantities of manure and fertiliser as in Denmark, Germany and the southern part of Sweden. A high rate of industrial activity coupled with a low degree of wastewater purification can also contribute to high area-specific nitrogen load. High area-specific phosphorus load is related to high population density (as e.g. in the Western Baltic and in the Sound), a high rate of industrial activity and to some extent to the intensity of agricultural activity. A low degree of wastewater purification also leads to higher area-specific nutrient load.

The area-specific load is lowered by processes such as retention, nutrient turnover as a denitrification processes and high groundwater influx to the rivers (see 5.3.3). These processes will generally affect large rivers, leading to lower area-specific nutrient load and lower flow-weighted concentrations compared with smaller rivers.

5.4 Heavy metals load going into the Baltic Sea

Riverine and direct load of heavy metals into the Baltic Sea is an environmental problem. The long-term effects of accumulation of some of the metals, e.g. mercury and cadmium in biota, are well known. The eventual fate of the metals when they finally come into contact with the bottom sediments is another cause for concern. Other processes such as eutrophication and anoxicity, may greatly influence the distribution of the metals.

The total pollution load of metals in marine waters varies among the different subregions depending on the population density, location of industries and the exploitation of natural resources. The anthropogenic load derives, for instance, from industrial wastewater, leakage from products in use and those removed from service, "natural" degradation of products and pollution from various types of land-use (for example fertilising: excess of cadmium in fertilisers) and mining (mine waste deposits).

Due to very incomplete data, a full picture of the heavy metals load, going into the Baltic Sea could not be given. The difficulties in obtaining comparable heavy metals load data in 1995 were caused by a lack of laboratory equipment for analysis, inability to ensure adequate sampling or difficulty in analysing very small concentrations of certain metals. Both countries in transition and other Contracting Parties encountered one or all of these problems. There are also very apparent differences in the size of the water courses in the various countries. For example, southern Sweden's relatively few large rivers could be sampled effectively, while the hundreds of small streams and brooks in Denmark were very difficult to sample in practice.

Therefore, many figures are missing from this heavy metals summary, but reported results indicate that riverine heavy metals load is the largest source of total pollution load with approximately 90% except perhaps for cadmium. The municipal and industrial wastewater discharges together with diffuse discharges within the river catchment areas are probably the main sources within the riverine load. The cadmium load data from Russia concerning municipalities and industrial plants discharging directly into the Gulf of Finland are very high. Although the figures for heavy metals load are very uncertain, Table 5.10 provides a summarised overview of the heavy metals pollution load on the whole Baltic Sea. In Tables 5.11 to 5.15 a more detailed overview of the heavy metals load by subregion is given, whereby no total load figures are presented due to the lack of many data.

Table 5.10: Heavy metals load going into the Baltic Sea from rivers, municipalities and industrial plants in 1995 (Data is incomplete.^{1,2,3})

¹ All figures are missing for Denmark.

 $^{\rm 2}$ Figures for rivers in Latvia (Hg only, because of lack of proper equipment for Hg) and for rivers in Russia (Hg and Cd only) are missing.

³ All Estonian figures are from 1994.

		Load from	Load from	Treated	Untreated	Treated	Untreated	TOTAL	TOTAL	AREA
Hg		monitored	unmonitored	municipal	municipal	industrial load	indurstrial	Hg	DRAINAGE	SPECIFIC
		rivers	rivers and coastal	load	load discharging directly into the Baltic Sea		load	LOAD	AREA	Hg LOAD
			areas							
		in kg/a	in kg/a	in kg/a	in kg/a	in kg/a	in kg/a	in kg/a	in km	in kg/km
	F1 SE	575 118	10 #	2,4 10		22 20		609	133 167 118710	0,0046
BOB	Sum	693	10	12		42		148 757	251877	0,0012 0,0030
	F1	94	$\#$	1,6		4,0		100	39 301	0,0025
	SE	144	$\#$	4,0		$\#$		148	170 088	0,0009
BOS	Sum	238		5,6		4,0		248	209389	0,0012
	F1	4,0	$\#$	6,6		$\#$		$\overline{11}$	8952	0,0012
ARC	Sum	4,0		6,6				$\mathbf{1}$	8952	0,0012
	F1	78	$\#$	9,7		$\#$		88	49 703	0,0018
	EE ¹⁾	886	$\#$	\ast	$\#$	\ast	$\#$	886	287 641	0,0031
	RU	\ast		548	$\#$	223	$\#$	771	67 357	0,0114
GUF	Sum	964		558		223		1745	404 701	0,0043
	EE ¹⁾	117	$\#$	$\#$	$\#$	$\#$	$\#$	117	17018	0,0069
	LV	\ast	$\#$	\ast	#	\ast	$*$		117941	
GUR	Sum	117						117	134959	0,0009
	EE ¹⁾			#	$\#$					
	L _V	\ast	$\#$	\ast		\ast			13 602	
	LT	#	$\#$	\ast	$\#$	#	$\#$		98 890	
	RU	\ast	$\#$	\ast	$\#$	\ast	$\#$		15 000	
	PL	9 2 7 0	$\#$	77	76	342	0,25	9765	331 196	0,0295
	DE	46		1,0		0,034	<u>—</u>	47	9668	0,0049
	DK	\ast	$\#$	\ast		\ast			1206	0,0000
	SE	62	$\#$	35		$\#$		97	67 766	0,0014
BAP	Sum	9378		113	76	342	0,3	9909	537328	0,0184
	DE	57	0,5	2,0		0,0024		60	7 3 5 9	0,0081
	DK	\ast	#	102		\ast		102	12342	0,0083
WEB	Sum	57	0,5	104		0,002		162	19701	0,0082
	DK	\ast	#	155		\ast		155	1737	0,0892
	SE	0,1	$\#$	6,0		0,1		$\pmb{6}$	2 4 0 9	0,0026
SOU	Sum	0, I		161		0,1		161	4 1 4 6	0,0389
	DK	\ast	$\#$	67		\ast		67	15826	0,0042
	SE	117	$\#$	41		\mathcal{L}		159	67 435	0,0024
KAT	Sum	117		108		\boldsymbol{l}		226	83 26 1	0,0027

Table 5.11 Riverine and direct point source Mercury load going into the Baltic Sea in 1995 by subregion

 $-$ = nothing to report (this source does not exist)

1) Estonian riverine Hg load is from 1994

 $* =$ data not available (should have been reported)

 $# =$ data not available, but not obligatory

Table 5.12 Riverine and direct point source Cadmium load going into the Baltic Sea in 1995 by subregion

 $-$ = nothing to report (this source does not exist)

1) Estonian riverine Cdload is from 1994

*=data not available (should have been reported)

 $\# =$ data not available, but not obligatory

		Load from	Load from	Treated	Untreated	Treated	Untreated	TOTAL	TOTAL	AREA
Zn		monitored	unmonitored		municipal municipal industrial		indurstrial	Zn	DRAINAGE	SPECIFIC
		rivers	rivers	load	load	load	load	LOAD	AREA	Zn load
			and coastal		discharging directly into the Baltic Sea					
			areas							
		in kg/a	$\frac{1}{2}$ in kg/a	in kg/a	$\frac{1}{2}$ in kg/a	in kg/a	in kg/a	in kg/a	in km	in kg/km
	F1	169 740	30 500	1640		7670		209 550	133 167	1,57
	SE	267 040	17 100	1090		2800		288 030	118710	2,43
BOB	Sum	436780	47600	2 7 3 0		10470		497580	251877	1,98
	F1	133 380	58 400	1070		37911		230 761	39 301	5,87
	SE	553 250	43 000	1756		1030		599 036	170 088	3,52
BOS	Sum	686 630	101400	2826		38941		829797	209389	3,96
	F1	25 / 20	50 200	2070		108		77 498	8952	8,66
ARC	Sum	25 120	50 200	2070		108		77498	8952	8,66
	F1	69 140	23 200	6 3 8 0		63		98 783	49 703	1,99
	EE ¹⁾	140 660	$\#$	\ast	$\#$	\ast	$\#$	140 660	287 641	0,49
	RU	581 147	4 190	260 745	#	33 845	#	879 927	67 357	13,06
GUF	Sum	790 947	27390	267 125		33 908		1119370	404 701	2,77
	EE'	19020	$\#$	$\#$	$\#$	$\#$	$\#$	19020	17018	1,12
	LV	96 230	4 0 8 0	2 7 5 3	4012	354	6,0	107 435	117941	0,91
GUR	Sum	115250	4080	2 7 5 3	4012	354	6,0	126455	134959	0,94
	EE			$\#$	#					
	LV	11920	980	54		78		13 032	13 602	0,96
	LT	89 751	$\#$	3 400	$\#$	$\#$	$\#$	93 151	98 890	0,94
	RU	190	$\#$	\ast	$\#$	\ast	$\#$	190	15 000	0,01
	PL	801 330	$\#$	11076	19923	3 5 7 2	72	835 973	331 196	2,52
	DE	8018		896		13		8927	9668	0,92
	DK	\ast	$\#$	\ast		\ast			1206	
	SE	119692	114 000	13 670		247	\equiv	247 609	67 766	3,65
BAP	Sum	1030901	114980	29 096 19 923		3910	72	1198882	537328	2,23
	DE	6 4 5 7	506	3 3 8 1		28		10372	7 3 5 9	1,41
	DK	\ast	$\#$	5 9 0 8		\ast		5 908	12342	0,48
WEB	Sum	6457	506	9289		28		16280	19701	0,83
	DK	\ast	$\#$	8722		\ast	$\overline{}$	8722	1737	5,02
	SE	601	7 000	1485		$\#$	$\overline{}$	9086	2 4 0 9	3,77
SOU	Sum	601	7000	10207				17808	4 1 4 6	4,30
	DK	\ast	$\#$	3 7 8 7		\ast		3 7 8 7	15826	0, 24
	SE	136 040	2 3 0 0	6838		129	$\qquad \qquad$	145 307	67 435	2,15
KAT	Sum	136 040	2 3 0 0	10625		129		149094	83 26 1	1,79

Table 5.13 Riverine and direct point source Zinc load going into the Baltic Sea in 1995 by subregion

 $-$ = nothing to report (this source does not exist)

1) Estonian riverina Zn load is from 1994

* = data not available (should have been reported)

 $# =$ data not available, but not obligatory

² data not complete

Table 5.14 Riverine and direct point source Copper load going into the Baltic Sea in 1995 by subregion

 $=$ nothing to report (this source does not exist)

1) Estonian riverine Cu load is from 1994

 $\text{*} =$ data not available (should have been reported)

 $\# =$ data not available, but not obligatory

Pb		Load from monitored rivers	Load from unmonitored rivers and coastal areas	Treated municipal load	load	Untreated Treated municipal indurstrial industrial load	Untreated load discharging directly into the Balrtic Sea	TOTAL Pb LOAD	TOTAL DRAINAGE AREA	AREA SPECIFIC Pb LOAD
		in kg/a	in kg/a	in kg/a	in kg/a	in kg/a	in kg/a	in kg/a	in km	in kg/km
	F1	8870	1600	74	$\qquad \qquad -$	9,0	$\overline{}$	10 553	133 167	0,079
	SE	5 0 7 4	$\#$	28	$\qquad \qquad \overline{\qquad \qquad }$	1440	$\qquad \qquad -$	6 5 4 2	118710	0,055
BOB	Sum	13944	1600	102	—	1449	$\overline{}$	17095	251877	0,068
	F1	8 9 8 0	1500	50	$\qquad \qquad \qquad$	1231		11761	39 301	0,299
	SE	10754	$\#$	56	$\overline{}$	50		10860	170 088	0,064
BOS	Sum	19734	1500	106	$\overline{}$	1281	$\overline{}$	22 621	209389	0,108
	F1	1430	2 900	110	$\overline{}$	75	—	4515	8952	0,504
ARC	Sum	1430	2 900	110		75		4515	8952	0,504
	F1	4 3 3 0	1650	290	÷,	34	-	6 3 0 4	49 703	0, 127
	EE ¹⁾	21 680	$\#$	\ast	$\#$	\ast	$\#$	21 680	287 641	0,075
	RU	63 090	190	25 825	$\#$	37	$\#$	89 142	67 357	1,323
GUF	Sum	89 100	1840	26 115		71		117126	404 701	0,289
	EE ¹⁾	4 6 4 0	$\#$	$\#$	$\#$	$\#$	$\#$	4 6 4 0	17018	0,273
	$L V$	6 2 2 0	260	409	477	147	\ast	7513	117941	0,064
GUR	Sum	10860	260	409	477	147		12 153	134959	0,090
	EE ¹⁾		$\qquad \qquad -$	$\#$	$\#$					
	LV	896	72	\ast	$\overline{}$	34	$\qquad \qquad -$	1002	13 602	0,074
	LT	17976	$\#$	\ast	$\#$	$\#$	$\#$	17976	98 890	0,182
	RU	☀	$\#$	\ast	$\#$	\ast	$\#$		15 000	
	PL	124 630	$\#$	2899	835	708	2,0	29 074	331 196	0,390
	DE DK	415 \ast	$\qquad \qquad -$ $\#$	211 \ast	$\overline{}$ —	16 \ast	$\qquad \qquad \longleftarrow$ $\overline{}$	642	9668 1206	0,066
	SE	4 2 1 2	$\#$	396	$\qquad \qquad \qquad$	114	$\qquad \qquad \qquad$	4 7 2 2	67 766	0,070
BAP	Sum	148 129	72	3 5 0 6	835	872	2,0	153 416	537328	0,286
	DE	969	84	351		1,0		1405	7 3 5 9	0,191
	DK	\ast	$\#$	163		\ast		163	12342	0,013
WEB	Sum	969	84	514		1,0		1568	19701	0,080
	DK	\ast	$\#$	198		\ast		198	1737	0, 114
	SE	25	#	40				65	2 4 0 9	0,027
SOU	Sum	25		238				263	4 1 4 6	0,063
	DK	\ast	$\#$	103		\ast		103	15826	0,007
	SE	8 0 5 2	$\#$	423		66		8 5 4 1	67 435	0, 127
KAT	Sum	8052		526		66		8644	83 26 1	0, 104

Table 5.15 Riverine and direct point source Lead load going into the Baltic Sea in 1995 by subregion

 $=$ nothing to report (this source does not exist)

1) Estonian riverine Pb load is from 1994

 $\text{*} =$ data not available (should have been reported)

 $\# =$ data not available, but not obligatory data not complete

6

Source apportionment

6.1 Introduction

Source apportionment is a tool for evaluating the importance of different sources to riverine nutrient fluxes. The objective of separating riverine fluxes is to assess the importance of anthropogenic sources. The political and administrative systems, therefore, could have a tool for evaluating what measures are the most cost effective for the environment in combating nutrient pollution. According to the PLC-3 Guidelines, the Contracting Parties should therefore estimate the proportion of natural (background) load and anthropogenic (point source and diffuse source) load.

Special Guidelines for the source apportionment were prepared at a workshop held in Denmark in June 1995, which are described in TC POLO 3/9, Annex 5 "Guidelines to estimate natural and anthropogenic contributions to riverine fluxes (source apportionment)". Three different methodologies for source apportionment were compared: the Danish, the German and the Finnish methodology. It was recommended that the Contracting Parties should use one of these three methodologies (TC POLO 3/9, Annex 5). In this chapter, the principles of source apportionment methodologies are described briefly, and the methodology used by each of the Contracting Parties is summarized.

6.1.1 Definitions

Anthropogenic nutrient sources are defined as load from human activities and can be divided into:

- a) Point sources (wastewater from municipal treatment plants, industrial plants, fish farms, stormwater reservoirs and other stormwater sewage constructions)
- b) Diffuse sources: i) agriculture ii) scattered dwellings

In some of the Contracting Parties loads from fish farming, stormwater reservoirs and other stormwater sewage constructions are not measured. In these countries, the diffuse sources

consist of loads from agriculture, scattered dwellings, fish farming and stormwater constructions.

Separating loads from agriculture and scattered dwellings is very difficult, because:

- 1) Assessment of the potential load from scattered dwellings is difficult as it depends on the potential production of wastewater and the equipment used to collect and treat the wastewater if any treatment is applied;
- 2) An estimation of the part of the potential load reaching from scattered dwellings is very uncertain as the wastewater can infiltrate the soil, can be connected to tiles, collected in different kinds of containers. Further, the distance from the scattered dwelling to the recipient is an important factor.

Therefore, it was recommended that diffuse sources should include load from agriculture, from scattered dwellings and also the natural load (background load).

Natural sources (background load) of riverine nutrient fluxes are assumed to characterize the conditions in a watershed that is unaffected by human activity. Finding such watersheds is almost impossible as the amount of atmospheric deposition caused by human activity has been increasing in recent decades. Thus, the best estimate for background load is found in small, sparsely populated catchment areas with low human activity, such as areas of natural forest areas and/or uncultivated areas.

Assessing the importance of atmospheric deposition on riverine fluxes is not possible. From a scientific point of view, giving figures on the proportion of atmospheric deposition reaching the freshwater environment and the sea via riverine loads would be irresponsible. Only atmospheric deposition on larger bodies of freshwater can be evaluated in a source apportionment calculation.

In general terms source apportionment is assessed by subtracting the natural (background) load (NL) and the

point source load (PL) from the riverine load (RL) and thus obtaining an estimate of the diffuse load (DL):

$$
DL = RL - NL - PL, \tag{1}
$$

The importance of the different source is then expressed as:

Proportion of $NL = (NL/RL)$. 100%(2) Proportion of $PL = (PL/RL) \cdot 100\%$ (3) Proportion of $DL = (DL/RL)$. 100% (4)

The anthropogenic load (AL) is:

$$
AL = PL + (DL - NL)
$$
 (5)

Nutrients from anthropogenic and natural sources are affected by temporary and more permanent sinks and by cyclical and removal processes (e.g. denitrification, retention in lakes and flooded riparian areas) which can be expressed as the retention of nutrients (RET). To assess the importance of different sources these processes must be taken into account as the measured riverine load only expresses the net riverine export. If retention is omitted from the source apportionment calculation, the diffuse load from agriculture and other sources will be underestimated. The source apportionment should therefore be based on the gross riverine load (GRL), i.e.:

$$
GRL = RL + RET \tag{6}
$$

and equations 1, 2, 3 and 4 are amended thus:

 $DL = RL + RET - NL - PL$ (7) Proportion of NL (NL/GRL).100% (8) Proportion of $P = (PL/GRL) .100\%$ (9) Proportion of $DL = (DL/GRL).100\%(10)$

There are several methods for obtaining values for the background load, the point source load, the diffuse load and the transport of nutrients. These methods range from measurements, empirical relations, emission coefficients or values based on experience.
6.1.2 Background load and load from unmonitored rivers and coastal zones

The background load is a figure for the natural load from a catchment area that is not affected by human activity. Finding any catchment area in the Convention area which fulfils this condition is not possible, as the atmospheric deposition today is higher than, for example, 100 years ago. The best obtainable background load figures are obtained by measuring the load from small catchment areas with natural unmanaged forests and/or catchment areas with very low agricultural and other human activities. The catchment areas used should therefore be sparsely populated.

Often a part of the catchment area is unmonitored and load estimations from these areas are necessary to perform a proper source apportionment. Load from unmonitored parts of rivers (e.g. coastal zones) or unmonitored rivers themselves can be estimated by more or less sophisticated methods. The most frequently used methods are:

- a) use of area coefficients for the diffuse load;
- b) use of discharge-weighted concentrations;
- c) use of models based on land use, climate, soil types etc.

a) Area coefficients:

Area runoff coefficients (kg N/ha and kg P/ha) may be determined from monitoring stations situated further upstream within the catchment area or from monitored catchment areas in which the soil type and land use corresponds to those of the unmonitored catchment areas. Before calculating the area coefficient the point source load must be deducted. The point source load for the unmeasured catchment areas are then added to the estimated diffuse load.

In some of the Contracting Parties no figures are available for the point source load in monitored or unmonitored areas. Therefore, a rough calculation is applied by extrapolating the figures from the monitored part of watersheds to correspond to the whole catchment area.

b) Discharge-weighted concentrations:

Use of discharge-weighted concentrations (the annual transport of a species of nutrient divided by the annual runoff of water) is recommended as the best estimate for the diffuse load from unmonitored areas. The optimum solution is to use measurements from agricultural catchment areas without point sources where soil type and land use correspond to those of the unmonitored catchment area. The discharge-weighted concentrations are then multiplied by measured or estimated runoff values. If there are no measured catchment areas without point sources with similar conditions to the unmonitored watershed, point source loads are deducted from the nutrient transport measured in a monitored watershed before the discharge-weighted concentration is calculated. Often an average of dischargeweighted concentrations from several measured catchment areas can be used for the unmonitored areas.

c) Models:

Empirical models can also be applied to estimate the diffuse load from unmonitored areas. The models can be based on relations between land use (proportion of cultivated areas, fertiliser and manure consumption, soil types, livestock etc.) in the watershed and the nutrient runoff to yield emission coefficients. These empirical models can be based on measurements combined with GIS information on land use, agricultural practices, soil types, livestock, use of manure and fertiliser etc.

6.1.3 Point source load and other methodological problems

Different approaches have been used to estimate the point source load:

- a) direct measurements or empirical load figures
- b) calculated load figures
- c) indirect estimations

In some of the Contracting Parties the point source loads are generally based on measurements, and estimations are only done for small point sources.

In some of the Contracting Parties measurements of the point source loads in the monitored part of their watershed are lacking. The point source loads are estimated from the number of inhabitants and volume of industry connected to municipal wastewater treatment plants multiplied by the values of nitrogen and phosphorous content in the potential load from one person equivalent. The load from point sources is sometimes determined from load estimations during low flow conditions. Under low flow conditions it is assumed that the increase in nutrient transport from a river monitoring station situated upstream of a town to a corresponding monitoring station situated downstream is tightly correlated to the point source load from the city. Although these estimated figures for point source loads are very uncertain, they are used in this chapter for some of the Contracting Parties to give an impression of the importance of different load sources.

Source apportionment is performed for one or more of the following categories:

- a) monitored rivers
- b) total riverine load to the sea excluding load from direct point sources
- c) total riverine load to the sea including load from direct point sources

The total riverine load is compiled on either a regional or on a nationwide scale. Performing a source apportionment for the total riverine load entering the Baltic Sea is not possible as information is incomplete or missing. Doing a source apportionment for all the Contracting Parties and all of the nine major regions of the Baltic Sea is also impossible. In chapter 6.3 the best obtainable results based on calculations and information of the Contracting Parties supplied with calculations made by the drafting group are shown. These results can only be compared with great caution and may be seen as a first attempt to give a rough estimate of load sources in the Contracting Parties.

6.2 Methodology used in the Contracting **Parties**

The applied source apportionment methodology in the Contracting Parties is briefly discussed in this chapter. The three recommended source apportionment methods can be divided into two groups.

a) the Danish and the Finish methods; b) the German immission method.

Methods in group a) are based on measurements of the nutrient transport in rivers and measurements or estimations of point source load and background load. It involves data from one year.

The German immission methodology is based on an analysis of relationships between measured concentrations and water flow in rivers, on flowload relationships and some information on larger point sources. It involves data from at least three or four years.

The Contracting Parties have applied one of the three methods with smaller or larger modifications as described below.

The Contracting Parties have estimated the background load from measurements in typical small unmanaged forested and/or non-cultivated catchment areas. The figures for background load therefore include not only natural background load and load from atmospheric deposition and from scattered dwellings, but also to some extent contributions from forestry and agricultural activity (if present).

6.2.1 Finland

The Finnish source apportionment methodology is described in detail in TC POLO 3/9, Annex 5). Finland performed the source apportionment for the total riverine load including direct point source load and load from unmonitored areas to the Bothnian Bay, the Bothnian Sea, the Archipelago Sea, the Gulf of Finland and totally for Finland in 1995. Source apportionment is not available

for monitored rivers. Source apportionment is determined according to equations (1) to (5).

Altogether 28 rivers were included in the Finnish monitoring programme in 1995 covering approximately 95% of the Finnish Baltic Sea catchment area. Riverine load from the remaining 5% of the catchment area was estimated for each subregion by extrapolation from the monitored part of the catchment area (area coefficients).

Point source load was compiled on the basis of the obligatory monitoring programmes run by municipal wastewater treatment plants and industrial plants. It is assumed that the point source loads entering rivers remain quantitatively unchanged in the rivers, i.e. no retention is accounted for, as only the lower catchment areas below larger lake basins were taken into account. In catchment areas with many lakes the retention of nutrients of anthropo-genic origin was assumed to be near 100%. Point source loads only include load from municipal wastewater treatment plants and industrial plants, which are given separately. Loads from scattered dwellings, stormwater reservoirs and other stormwater constructions and from fish farming are included in the diffuse load.

Background load was estimated from studies in forested catchment areas and is:

- a) $10\,\mathrm{kg}$ P/(km $^2\cdot$ a) and 250 kg N/(km $^2\cdot$ a), in the catchment areas of the Gulf of Finland, the Archipelago Sea and the Bothnian Sea;
- b) 10 kg P/(km² · a) and 170 kg N/(km² ·a), in the catchment area of the Bothnian Bay.

Due to retention in the catchment area the background load of phosphorus was estimated to decrease by 20 % in the three southern catchment areas (Gulf of Finland, Archipelago Sea, and Bothnian Sea) and by 25 % in the catchment area of the Bothnian Bay. For nitrogen, the retention percentages were estimated to be 15 % and 20 %, respectively. The background load figures included, to some extent, contributions from scattered dwellings etc.

The diffuse load is estimated by subtracting background load and point source load from the measured riverine transport (equation 1).

With the Finnish methodology, retention is only partly included and the anthropogenic load is therefore underestimated.

6.2.2 Russia

Russia has, in principle, used the Finnish methodology for source apportionment (equations 1 to 5). Russia have given information for 18 rivers, but as information on either point source load, background load or diffuse load (or total measured riverine load) is missing, the source apportionment has only been done for 8 or 9 rivers. Source apportionment is missing for the total Russian load entering the Gulf of Finland and Baltic Proper and for the total Russian riverine load. The source apportionment has not been done for the River Narva as information from Russia is missing.

Point sources as a rule include load from municipal wastewater treatment plants and industrial plants, and dividing the load from these sources is not possible. It seems that the phosphorus load from point sources is calculated on the basis of emission factors for urban areas (90 kg $P/(km^2 a)$). No figures are given for nitrogen. Background load was estimated from measurements in small non-agricultural catchment areas and forested areas yielding:

- a) 7 kg $P/(km^2 \cdot a)$ for the northern coast of the Gulf of Finland;
- b) 6 kg $P/(km^2 \cdot a)$ for the southern coast of the Gulf of Finland.

No figures are given for nitrogen. Determining whether the background load of phosphorus is deducted from the diffuse load in agricultural catchment areas is not possible.

The agricultural load that also includes load from scattered dwellings, smaller settlements and stormwater constructions is calculated by deducting the point source load and background load from the measured riverine load (equation 1). For unmonitored rivers in the coastal zone of the Gulf

of Finland an area coefficient of 15 kg P/km2 is used. Retention is not included in the Russian methodology.

The method gives only a very rough estimate of the importance of anthropogenic sources.

6.2.3 Estonia

Estonia has in principle used the Finnish methodology for source apportionment (equations 1 to 5). Source apportionment has been done for 14 monitored rivers, but not for the total Estonian load entering the Gulf of Finland, the Gulf of Riga or the Baltic Proper or for the total Estonian riverine load. Source apportionment has not been done for the River Narva as information from Russia is missing.

Point sources include load from municipal wastewater treatment plants and industrial plants, and dividing the load from these sources is not possible. No information is given concerning the way in which the point source load is obtained (i.e. whether measured or calculated /estimated).

The background load was estimated from measurements in small nonagricultural catchment areas yielding:

- a) $430 \text{ kg} \text{ N/(km}^2 \cdot a)$;
- b) $12 \text{ kg } P / (\text{km}^2 \cdot \text{a})$.

The load from forested areas is included as background load, since Estonian forests as a rule are managed without fertiliser consumption. It appears that background load is not deducted from the diffuse load in agricultural areas.

Agricultural load that also includes load from scattered dwellings, smaller settlements and stormwater constructions is calculated by deducting the point source load and background load from the measured riverine load (equation 1). Retention is not included in the Estonian methodology.

The method gives only a very rough estimate of the importance of anthropogenic sources.

6.2.4 Latvia

Latvia has in principle used the Danish methodology for source apportionment but without retention. Source apportionment has been done for 8 monitored rivers. Further, the author of this chapter used the Latvian methodology and submitted data to perform source apportionment for the total riverine load including the load from unmonitored areas and the load from direct point sources entering the Gulf of Riga and the Baltic Proper and for the total Latvian riverine load entering the Baltic Sea. Latvia has deducted the riverine load from neighbouring countries from the total riverine load to estimate the source apportionment from Latvian sources. The diffuse load from unmonitored areas is calculated on the basis of area coefficients from corresponding managed and non-managed areas.

The anthropogenic load is calculated by deducting the background load from the total riverine load. The diffuse load was calculated by deducting the load from point sources from the anthropogenic load.

The load from point sources in the diffuse coastal zone is known, but the way the load from point sources for monitored river catchment areas is obtained is not explained in detail. The number of inhabitants living in rural scattered settlements was estimated by deducting the number of inhabitants living in rural settlements from the number of inhabitants in administrative subregions. The point source load is assumed to include load from municipal wastewater treatment plants and from industrial plants.

The background load was estimated from measurements in small nonmanaged natural watersheds included forested areas yielding:

a) 750 (50-1 000) kg N/(km²·a); b) 35 (10-60) kg $P/(km^2 \cdot a)$

These are quite high background values, which assume the inclusion of loads from scattered dwellings, small towns and some industrial plants. Nonmanaged and managed areas and areas with rural scattered settlements were

estimated from land use data. The load from forested areas is not deducted from the background load. Further, the background load is not deducted from the diffuse load in managed areas.

6.2.5 Lithuania

Lithuania has used a combination of the Danish and the German source apportionment methodology.

Source apportionment has been done for two rivers (the Nemunas and the sum of the rivers Akmena-Dané and Sventoji). Further, the author of this chapter used the Lithuanian methodology and submitted data to perform source apportionment for the total Lithuanian riverine load including the load from unmonitored areas and the load from direct point sources entering the Baltic Sea (Baltic Proper). Lithuania has deducted the riverine load from neighbouring countries from the total riverine load to estimate the source apportionment for Lithuanian sources only.

The background load was estimated from measurements in three watersheds with low agricultural activity (percentage of arable land less than 20%) yielding:

- a) 0.32 to 0.80 mg N/l ;
- b) 0.05 to 0.09 mg P/l.

The extension of arable land and forested land was calculated from land use maps and cadastral registers. No information is given on whether the background load has been deducted for the total river catchment area.

Measurements in 5 rivers draining watersheds with more than 50% arable land were used to estimate area coefficients for unmonitored agricultural watersheds and to estimate the diffuse transport from agricultural land.

The load from point sources is not measured. The load from point sources is estimated by deducting natural background load from the anthropogenic diffuse transport i.e. the background load and the load from agricultural land were deducted from the gross riverine transport. Distinguishing between the load from municipal wastewater treatment plants and from indus-

trial plants is not possible. In the agricultural load the point source load is included to some unknown extent.

Retention was calculated using Bhrendt's formula (TC POLO 3/9, Annex 5):

$$
RET_N = 41.456 * q^{1.297} * C_N^{-0.542}
$$
 (11)

$$
RET_P = 28.13 q^{1.708},
$$
 (12)

where q is the specific runoff in l/(s . km^{2}) and C_N is the average concentration of dissolved inorganic nitrogen in the river concerned.

The gross riverine load of nitrogen and phosphorus is also calculated using Behrendt's formula as:

Net load nitrogen/gross load nitrogen $= 1 / (1 + RET_{N})$ 13) Net load phosphorus/gross load phosphorus = $1/(1 + RET_p)$ (14)

Behrendt's retention formulas (11 and 12) are set up for some German rivers and the Danish River Gudenaa, but have not been set up for, e.g. Lithuanian rivers. Retention is not included in the background and agricultural load area coefficients used for unmonitored areas as the retention is assumed unimportant in the small watershed from which these area coefficients are determined.

6.2.6 Poland

The Polish source apportionment methodology is based on calculated emission coefficients for background load and load from agriculture. Source apportionment is done for 12 rivers. Further, the author of this chapter based on submitted data performed source apportionment for the total riverine load including load from unmonitored areas (coastal zones) from Poland to the Baltic Sea (Baltic Proper).

25 rivers were monitored within a 2 to 6-year cycle. These comprise 41 small agricultural watersheds and 5 forested watersheds both without any major point sources (municipal wastewater treatment plants and industrial plants). The resulting data were representative of 4 major physical-geographical regions of Poland. The following parameters were included in the modelling: morphology, soil permeability, fertiliser consumption and hydrology. The unit emission of nitrogen and phosphorus from the 5 experimental forested watersheds were used as background values and deducted from the unit values measured in the experimental agricultural watershed to obtain emission values from agriculture. Unit loads from agriculture were corrected only in relation to the amount of arable land in the watersheds. From this information the unit load was divided into background emission and agricultural emission. On the basis of data from 2 to 6 years, unit loads were recalculated to consider hydrological condition. In the calculations it is assumed that the amount of emitted nutrients is proportional to riverine runoff. The unit loads for nitrogen were calculated as:

$$
Lr = (0.31536 * Cwr * Q * Z1993) / Zexp
$$
(15),

where:

- L_r is a nitrogen (phosphorus) unit outflow from agriculture in kg/(ha · a);
- Cw_r is the discharge-weighted nitrogen (phosphorus) concentration from agriculture in the runoff in mg/l;
- Q is the mean runoff during the study period in $1/(s \cdot km2)$;
- Z_{1993} is the fertiliser consumption (divided into mineral and natural) for the given area in 1993 in kg/(ha · a);
- Z_{exp} is mean fertiliser consumption (mineral and natural) in the catchment areas in kg/(ha . a);

$$
L_t = 0.31536 * Cw_t * Q \tag{16}
$$

where:

- L_{i} is the background nitrogen (phosphorus) unit outflow in $kg/(\text{ha} \cdot \text{a})$;
- Cw_t is the discharge-weighted nitrogen phosphorus) concentration in mg/l
- Q is the mean runoff during the study period for the area concerned in l/(s . km²).

These corrected unit loads were average for the physical-geographical regions for four soil types, as dischargeweighted concentrations of nitrogen and

phosphorus were calculated for each soil type in relation to runoff from fields and mean nutrient input from mineral and natural fertilisation during the study period. This procedure provides input data for a spatial model of the nitrogen outflow for the entire territory of Poland that has been set up for administrative units and for catchment areas. Data on soil types, morphology, hydrology and proportion of arable land is available on GIS.

Further, nitrate emissions from point sources were estimated, but no information is given about the methodology, nor about the methodology for phosphorus emissions from point sources. The emission coefficients from background, agriculture and point source loads then made it possible to calculate the importance of each source to the measured riverine load.

The ratio between the measured riverine load and the sum of the calculated emissions is an estimate of the retention within the watershed, the lakes and riparian areas.

6.2.7 Germany

Germany has applied two methods of source apportionment: the immission and the emission method. Germany performed the source apportionment for the total riverine load including direct point source load and load from unmonitored areas. Further, the source apportionment was done for 12 rivers (see chapter 6.3). For 9 of the 12 rivers investigated the emission method was applied and for the remaining 3 rivers the immission method was applied.

Both applied methods are quite different from the Finnish and Danish methodology. The emission method is fully independent of the observed load and uses a point source inventory and estimations on the different diffuse pathways via groundwater, interflow and surface runoff. Further atmospheric deposition, emissions from urban areas, erosion and agricultural direct emissions were considered. The diffuse emissions were estimated using GIS-data on landcover, soil types, elevation and statistical data on drainage areas, livestock and fertilizer consumption.

For three of the 12 investigated river basins the immission method was applied, where the separation of point

and diffuse sources was derived from the discharge dependency of the observed concentrations and loads. The total emissions of these three rivers were estimated by the application of the Behrendt retention formula (1996).

Because both German methods do not implicitly include background emissions, the proportion of the background load of the total load was estimated based on the following assumptions:

- Emissions from the diffuse sources are zero with exception of the emission by groundwater (base flow);
- Flow is only caused by base flow;
- The nutrient background concentrations are estimated as the mean nutrient concentrations in groundwater measured in a distant groundwater table.

The background load was estimated using concentrations of 2 mg N/l and 0.025 mg P/l, and multiplying these values with the corresponding flow in the rivers.

Point sources are municipalities and industry. The method assumes that point source loads are relatively constant and only slightly related to meteorological factors. Further, it is assumed that point sources are evenly distributed within the catchment.

Retention is not directly incorporated in the German methods, but the retention can be estimated from the comparison of the results of the emission method with the observed load (for the emission method) and of the point emissions and the portion of point load at the observed load derived by the immission method. The results of these comparisons support the empirical formula of Behrendt (1996), which shows that the normalized retention is increasing with decreasing area-specific runoff. It is assumed that processes such as denitrification and sedimentation equally affect the load from different sources.

These formulas give the ratio between nutrient load and nutrient emission, where the load is measured in the rivers. From these empirical equations it is also possible to estimate retention.

Anthropogenic sources are calculated by deducting the calculated background load and the load from pointsources from the total riverine load.

6.2.8 Denmark

The Danish source apportionment methodology is also described in detail in TC POLO 3/9, Annex 5. Denmark performed the source apportionment for the total riverine load including direct point source load and load from unmonitored areas for the Danish part of the Baltic Proper, the Western Baltic, the Sound, the Kattegat and for the whole of the Danish Baltic Sea catchment area. Further, the source apportionment was done for 103 streams and rivers, but these results are only given as pooled information in chapter 6.3 (Tables 6.3 and 6.4). Source apportionment is determined according to equations 5 to 10.

The 103 rivers and streams included in the Danish river monitoring programme for the compilation of riverine Baltic Sea pollution load correspond to a 50% to 60% coverage of the corresponding Danish Baltic Sea catchment area. Riverine loads from the remaining 40% to 50% unmonitored areas (including coastal areas) are estimated using either area coefficients or discharge-weighted concentrations multiplied by representative flow.

Point source loads from industrial plants and municipal wastewater treatment plants are measured for all plants bigger than 30 PE in monitored and unmonitored areas. Loads from small municipal wastewater treatment plants (a potential load less than 30 PE) are calculated in monitored and unmonitored areas. Loads from stormwater reservoirs and other stormwater constructions in monitored and unmonitored areas are calculated by means of precipitation and empirical equations. Further, loads from freshwater fish farms are calculated from food consumption and production, and from water flow and concentration measurements upstream and downstream of the fish farm. In Denmark point sources include loads from municipal wastewater treatment plants, industrial plants, stormwater constructions and fish farms so that making a real point source inventory is possible.

The potential load from scattered dwellings is calculated for monitored and unmonitored areas based on inhabitants living in scattered dwellings and the content of phosphorus (1 kg/a) and nitrogen (4.4 kg/a) per person equivalent (1 PE). On average, it is assumed

that approximately 50% of the potential load reaches streams and rivers, but this figure is uncertain. Load from scattered dwellings is a part of the diffuse load.

Background load is determined from measurements in 9 small (1-10 km2) scarcely populated catchment areas with low agricultural and forestry activity. The loads from these catchment areas in 1995 were:

- a) 7.3 kg P/(km² a) or a dischargeweighted concentration of
- 0.055 mg P/l (median value) b) 290 kg $N/(km^2 \cdot a)$ or a discharge-
- weighted concentration of 1.4 mg N/l (median value).

These background load figures are used for 94% of the Danish catchment area draining to the Baltic Sea assuming that the remaining 6% is fortified area without any background load.

Retention is included in the Danish source apportionment figures. Retention is calculated on the basis of intensive mass balance studies for 37 lakes. Results from these lakes and from calculation of the median retention coefficient in other lakes, combined with the percentage of lakes in the catchment area are used to calculate retention in monitored and unmonitored catchment areas. On average, nitrogen retention in Danish lakes was 25% of the input to the lakes in 1995. The corresponding figure for phosphorus was 3% .

Diffuse load was estimated by equation 7, where the gross riverine load is calculated as retention added to the measured riverine load. Agricultural load is calculated by deducting the load reaching streams and rivers from the diffuse load. Anthropogenic load is estimated by equation 5.

6.2.9 Sweden

Sweden performed the source apportionment for the total riverine load of nitrogen including direct point source load and load from unmonitored areas entering the Baltic Proper, the Sound and the Kattegat. No figures were given for phosphorus sources. Source apportionment is not available for the total Swedish nitrogen load entering the Baltic Sea. Further, source apportionment for nitrogen has been done for 115 subcatchment areas including unmonitored

areas. The Swedish methodology is an emission method, based on measurements and/or calculations of the loading sources (point sources and diffuse sources), and the total riverine load is the sum of these sources. The calculated riverine load is not comparable with the measured riverine load. The methodology is described in detail in Arheimer et.al. (1997).

Load from coastal point sources (i.e. municipal wastewater treatment plants and industrial plants) is calculated from measurements of nutrient concentration and flow in the effluents. The remaining load from point sources is measured or estimated.

Load (emission) from small settlements and scattered dwellings is calculated from statistics on population and wastewater production per capita. These emissions are included in the calculated diffuse emission. The diffuse emission also includes inputs from agricultural land and background load.

The total emission, or the gross riverine transport, is calculated with a nitrogen model. In this model some biological and chemical processes within the catchment area, the lakes and watercourses is also incorporated to estimate nitrogen turnover and retention. Nitrogen concentrations leaving the root zone from cultivated areas with various land use and three different soil types are calculated with the SOIL-N model. Information on land use and load from point sources is presented by using GIS tools for 3 725 sub-catchment areas in the Baltic Proper, the Sound and the Kattegat. Emissions from forested areas and noncultivated areas (natural areas) are estimated from measurements in some minor experimental catchment areas. Atmospheric deposition on lakes is calculated with the MATCH model.

Background emission is measured to amount to $200 - 700$ kg N/(km² a). The model is calibrated against concentration and flow measurements in streams and against measurements in experimental small watersheds. Retention within the catchment areas, in lakes and watercourses constitutes from less than 25% of the emission of nitrogen in coastal zones to more than 75% in subcatchment areas with high percentages of lakes.

The importance of different sources is calculated as the measured/ calculated emission divided by the calculated riverine load. The resulting source apportionment is to some extent an average for several years and not only based on figures from 1995.

6.3 Results of source apportionment

The performed source apportionments are given inTables 6.1 to 6.4. When comparing the figures it is important to take into account the different source apportionment methodologies and assumptions. Further, there is much missing data and few countries have done a point source inventory. Therefore, only very rough conclusion can be drawn.

Tables 6.1 and 6.2 suggests that anthropogenic sources are the most important sources of riverine nitrogen and phosphorus loads including direct point source loads in the Contracting Parties except Finland, where the background load has the same importance as anthropogenic sources. Further, in Lithuania the background load of phosphorus makes up approximately 50% of the total load, which compared with the corresponding nitrogen figure of 5% of the total nitrogen load is very high quota. In Finland and probably Sweden, the background load is most important in the catchment area within the northern part of the country (catchment areas of the Bothnian Bay and the Bothnian Sea). In fact in the catchment areas of the Archipelago Sea and the Baltic Proper, the anthropogenic sources are also the most important in Finland. This reflects the fact that with higher human activity, such as agriculture, industry and the number of inhabitants per km $^{\scriptscriptstyle 2}$ the anthropogenic load will rise in absolute figures and in relative importance.

Tables 6.1 and 6.2 also suggest the importance of point sources. In Latvia, Poland, Germany and Denmark, point sources make up a higher percentage of total riverine phosphorus load than the corresponding figures for nitrogen. In Poland, the load from point sources appears to have the highest importance of all the Contracting Parties. In Finland and Lithuania the opposite might be true, but it is important to remember that the load from point sources in many countries is very uncertain and that point sources do not include the same elements.

In regions with high population density and high industrial activ-ity, such as in the catchment area of the Sound, point sources are a very important phosphorus source.

The major element in the load from diffuse sources is often agricultural load. The diffuse load might be the most important source of total riverine nitrogen load, including load from coastal areas, in the catchment area of the Archipelago Sea in Finland, in Latvia, in Lithuania, in Germany, in Denmark and in the southern part of Sweden (Table 6.1). In areas with very intensive agriculture such as in Germany, Denmark and the southern part of Sweden the diffuse nitrogen sources make up approximately 80% of the total waterborne land-based load entering the Baltic Sea.

The diffuse phosphorus load might be the most important source of the total waterborne land-based load in Finland entering the Archipelago Sea, in Latvia the Gulf of Riga, in Germany the Baltic Proper and the Western Baltic and in Denmark the Kattegat (Table 6.2).

The source apportionment of nitrogen and phosphorus for the monitored rivers and streams (Tables 6.3 and 6.4) shows overall the same picture as the information in tables $6.\overline{1}$ and $6.\overline{2}$. There are of course greater differences as some monitored rivers are draining small, essentially natural areas, and therefore the background load is very important. Only the mean source apportionment is given for the Danish rivers and streams draining to the Baltic Proper, the Western Baltic, the Sound and the Kattegat, although source apportionment was done for the 103 monitored rivers and streams. The importance of the different nitrogen load sources on these Danish rivers and streams show only minor variations between the rivers and streams running into the Baltic Proper, the Western Baltic and the Kattegat, but high variation between those running into the Sound, as some streams drain very intensive populated areas with a high point source load. The variation of the importance of phosphorus load sources is higher for the Danish rivers and streams draining into the Baltic Proper, the Western Baltic, the Sound and the Kattegat, than for the corresponding nitrogen figures.

Table 6.2 Source apportionment for the total riverine phosphorus load including load from coastal areas given for the 9 Contracting Parties and their subregion catchment areas in 1995. Other point sources are load from stormwater construction and from freshwater fish farm.

Table 6.3 Source apportionment for the nitrogen load in monitored rivers in the 9 Contracting Parties in 1995. Other point sources are load from stormwater construction and from freshwater fish farms.

Table 6.4 Source apportionment for the phosphorus load in monitored rivers in the 9 Contracting Parties in 1995. Other point sources are load from stormwater construction and from freshwater fish farms.

Results, Conclusions and Recommendations

7

7.1 Results

The Third Baltic Sea Pollution Load Compilation (PLC-3) carried out in 1995 examined nutrient, organic matter and heavy metal loads entering the Baltic Sea marine environment from rivers, coastal areas and direct point sources such as municipalities and industrial plants. However, point source discharge was only taken into account when the source was discharging directly into the Baltic Sea or when it was located downstream of the hydrological/hydrochemical station in the river for which the load is given. Therefore, only a small proportion of the total point sources located within a river basin were considered, which means that an inventory of all point sources in the whole Baltic Sea catchment area is missing. The calculations in PLC-3 can be considered more reliable and precise than calculations in PLC-1 and PLC-2, but still many uncertainties remain due to incomplete data sets, especially from the Russian Baltic Sea catchment area and from nearly all the Contracting Parties where heavy metals are concerned.

The runoff and direct point source discharge considered in PLC-3 is 466000 million m $\frac{7}{a}$, of which about 99% represents runoff into the Baltic Sea from rivers and coastal areas. Only 1% of the discharge came from municipalities and industrial plants located at the coast, to which treated and untreated municipal discharge and treated industrial discharge each contributed about 0.3%. The amount of untreated direct industrial discharge was negligible in comparison with all other pollution sources.

In 1995 the total $\mathop\mathrm{BOD}\nolimits_7$ load going into the Baltic Sea amounted to 1 140 000 t. The major part of organic matter load, 80%, entered the Baltic Sea via rivers and coastal areas. The BOD₇ load from municipalities and industrial plants discharging treated wastewater directly into the Baltic Sea is 9% in each case. The share of untreated municipal BOD_7 load was quite low, only 1% , but it should be noted that the untreated portion of the load from the Russian Kaliningrad region discharging directly into the Baltic Proper is missing. The share of untreated direct industrial BOD₇ load was also quite low, 0.002%.

In 1995 the **total waterborne** N_{total} $input$ and P_{total} input (natural background and anthropogenic) into the Baltic Sea amounted to 761 000 t/a and 38 000 t/a, respectively. The major part of N_{total} load, 90% , and \overline{P}_{total} load, 80% , entered the Baltic Sea via rivers. The treated municipal and industrial share of N_{total} load discharging directly into the Baltic Sea comprised 8% and 2% respectively, whereas these shares of P_{total} load were 13% and 5%, respectively. The proportion of untreated municipal and industrial N_{total} load and P_{total} load discharging directly into the Baltic Sea was quite low, less than 0.3% in each case.

Due to very incomplete data, a full picture of the heavy metals load going into the Baltic Sea could not be given. Many figures are missing from the heavy metals summary, but reported results indicate that riverine heavy metals load is the largest source of total pollution load with approximately 90%. The municipal and industrial wastewater discharges, together with diffuse discharges within the river catchment areas, are probably the main anthropogenic sources within the riverine load.

Because much of the pollution load is introduced into the Baltic Sea via rivers, another important step forward was to distinguish the nutrient load between the natural (background) and anthropogenic contributions (point source and diffuse source load) to riverine fluxes. The so called source apportionment is a tool for evaluating the importance of different sources to riverine nutrient fluxes. The objective of separating riverine fluxes is to assess the importance of anthropogenic sources. The political and administrative systems, therefore could have a tool for evaluating what measures for combating nutrient pollution yield the best cost benefit for the environment. According to PLC-3 Guidelines, the Contracting Parties should therefore estimate the proportion of natural load and the anthropogenic load, separately.

According to the results of the source apportionment the natural background load of nitrogen contributed 10- 20% of the total load in Denmark, Germany and southern parts of Sweden, but more than 50% of the corresponding load in Finland. Diffuse load, mainly from agriculture, was the main nitrogen source in many Contracting Parties, between 70% and 90% of the total nitrogen load, but in Finland and Poland it was only one-quarter and one-third, respectively. The point source load is usually the minor source of nitrogen load constituting 15-20% of the total nitrogen load in Denmark, Germany and southern Sweden and up to approximately 40% in Lithuania and Poland. On the other hand, point sources were the largest source of phosphorus load in Denmark, southern Sweden, Germany, Poland and Latvia (50-65%), but the smallest in Finland (15%). Background load of phosphorus made up 10-15% in Denmark, Germany, Poland, Lithuania and southern Sweden and up to approximately 50% in Finland and Latvia. Diffuse sources constituted 20-40% of total phosphorus load. It is very important to mention that figures provided by the Contracting Parties were in some cases very uncertain. Moreover, no information was received from Russia or Estonia, so that the results given above should be used with caution. Within the Contracting Parties there is great variation in the importance of the load from different sources, depending on land use, soil types, percentage of agricultural land and population density.

7.2 Discussion of results and proposals for the next stage of PLC

During the third stage of PLC the major uncertainties and weaknesses of PLC-2 could be avoided by establishing a quality assurance system, creating a dataentry system closely connected to a database and by doing the first steps for a source apportionment. Compared with the former two Pollution Load Compilations PLC-3 is a significant step forward as it gives much more reliable and complete data on the total loads entering the Baltic Sea including the description of the methodologies used in the various countries. Compared with PLC-2, the coverage of the pollution sources in PLC-3 increased significantly as load from small rivers ($Q < 5$ m \sqrt{s}) and small settlements (< 10 000 PE) were included. Nevertheless, during PLC-3 many uncertainties remain due to incomplete data sets from Russia and from nearly all Contracting Parties concerning heavy metals. Because the quality of the reported load data during these three stages was very variable, it was still impossible to compare PLC-3 with former pollution load compilations, to evaluate trends in loads and to evaluate the importance of various pollution sources.

The aim of the discussion is to reveal some issues that have arisen in the course of the work and to draw attention to some problems that need to be solved before the work starts on the next Pollution Load Compilation (PLC-4). In the following the main shortcomings are discussed.

Information about the Baltic Sea catchment area

There is an essential need to check and review the information by subregion and by Contracting Party with regard to the Baltic Sea river catchment areas and coastal areas including information on how the population is settled. Especially for transboundary rivers there were many difficulties in dividing the catchment area and the population between the countries concerned. When a part of a transboundary river catchment area is in the territory of a Non-Contracting Party, it was very difficult to obtain any information about these shares, including population. Therefore, separation of riverine load into Contracting Parties and Non-Contracting Parties could not be done. For the next Pollution Load Compilation, it is very important that information (area, population, point sources and agriculture) should be given for each transboundary river catchment area as a whole and separately for each country's share (Contracting Parties as well as Non-Contracting Parties). Moreover, each Contracting Party should present information (area, population, point sources and agriculture) about the monitored and unmonitored parts of the river basins as well as coastal zones.

Rivers

Flow measurement is a key element when calculating riverine load. The river flow in large rivers varies in most cases more than the concentration. Thus, it is of great importance that the flow is registered continuously. The main rivers in the Baltic Sea catchment area have permanent hydrological stations corresponding to the WMO Guide to Hydrological Practice. The partially monitored rivers have no permanent hydrological stations. Partially monitored rivers should be avoided by making contemporary runoff measurements at the hydrochemical monitoring measurement site in order to obtain proper runoff information.

In most of the Baltic Sea rivers the permanent hydrological/hydrochemical stations are not located at the river mouth, as far as the eastern and southern coasts are concerned these stations are very often situated far from the coast, sometimes at a distance of 50-60 km. Since the way of submitting the riverine load for the unmonitored portion of monitored rivers was not clearly defined in the PLC-3 Guidelines, different approaches were used: in some countries, e.g. Finland, the load for the unmonitored portion of monitored rivers was calculated on the basis of similarity and added to the measured portion of monitored rivers, so that a total riverine load figure could be given. In other countries, e.g. Denmark, the load for the unmonitored portion of monitored rivers was also calculated on the basis of similarity, but presented under the source category, unmonitored rivers. A third approach used in all other Contracting Parties was that the unmonitored portion of the monitored rivers was measured and reported under the source category "point sources". In the next Pollution Load Compilation, a harmonised reporting procedure of the load for the unmonitored portion of monitored rivers should be precisely defined.

In comparison with PLC-2, the sampling frequency in the monitored rivers has been increased to at least 12 times per year for organic matter and nutrients. Only in Polish rivers the sampling frequency for all parameters measured was once per week. For heavy metals a sampling frequency of less than

12 times per year was mostly insufficient. In a lot of rivers no heavy metal measurements were carried out. For those rivers carrying the heaviest pollution load into the Baltic Sea such as the Neva, the Daugava, the Vistula, the Nemunas etc., the sampling frequency should be increased to once a week; the sampling procedure should take into account flow proportional sampling within the cross section.

Point sources (municipalities and industrial plants)

With reference to PLC-2 the number of industrial plants and municipal wastewater treatment plants with continuously measured and registered flow measurements has increased considerably. Although the portion of the untreated wastewater is lower in PLC-3 than in PLC-2 the main difficulties are connected with the measurement of untreated wastewater from smaller settlements, scattered dwellings, overflows, by-passes and other stormwater constructions. Measurements of untreated wastewater is a problem for countries in transition, especially for large cities such as Saint Petersburg and Riga, which have a large number of outlets. For the next PLC the measurement of untreated wastewater from all sources discharging directly into the Baltic Sea should be improved, so that all Contracting Parties provide the required information about loads, inhabitants, population equivalents and reduction coefficients for scattered dwellings.

As indicated in chapter 3, only in Denmark, Finland, Germany and Sweden the treated wastewater amount is measured continuously with an accuracy of more than 5%. In all the other Contracting Parties there is sometimes a high percentage of point sources for which the wastewater amount has not been measured. This was determined on the basis of water consumption and is therefore only an estimate. For the next PLC, the target should be to measure the amount of treated wastewater continuously in at least all large point sources (e.g. municipal wastewater treatment plants > 10000 PE).

Reliability of the results of PLC-3

The Contracting Parties are responsible for the quality and reliability of data. The main problem in PLC-3 was that not all obligatory parameters were measured by each Contracting Party. There was a big problem connected with Russian data because a lot of the figures were only estimated as totals by subregion, and figures from the Kaliningrad region were missing completely. Compared with PLC-2, heavy metal load data has improved a little, but there are still a lot of load figures missing, so that it is not possible to present total heavy metal load figures for the whole of the Baltic Sea.

The difficulties in obtaining comparable load data in 1995 were caused by a lack of laboratory equipment for analysis, inability to ensure adequate sampling or difficulty in analysing very small concentrations of certain substances. There are also very clear differences in the size of the water courses in the various countries. For example, southern Sweden's relatively few large rivers could be sampled effectively, while the hundreds of small streams and brooks in Denmark were very difficult to sample in practice.

In addition, the natural differences between the Contracting Parties and the consequences of different legislation should also be recognised. Due to the fact that Polish legislation does not allow load figures for industrial plants to be supplied plant by plant, Poland only submitted summarised load figures by branch of industry. Furthermore, some of the Contracting Parties monitored load from small plants with only 30 PE in monitored and unmonitored catchments, while others did even not monitor large municipalities.

The main task for the next PLC is that each of the Contracting Parties should report reliable and complete data sets of the pollution load. It will then be possible to estimate the total pollution load going into the Baltic Sea. Therefore, the parameters for which total loads of all source categories should be given must be obligatory in each source category, without footnotes and exceptions. It is useless to declare a parameter as being voluntary. All parameters must be mandatory.

Chemical Analyses and Quality Assurance

A quality assurance programme was established before starting PLC-3 to obtain reliable and comparable data. The first step was to establish national reference laboratories, which have played an essential role in improving analytical performance in most countries. The national reference laboratories have provided personnel training and carried out interlaboratory comparisons. Most laboratories that participated in PLC-3 carried out internal EN 45001 and ISO/IEC Guide 25. While PLC-3 was underway, some improvements were made, particularly in the reference laboratories of Estonia, Latvia, Lithuania and Poland with the support of the EU-PHARE programme. Unfortunately some countries had problems with, for instance, inappropriate methodology or inadequate instruments. However, the quality assurance programme has to some degree supported the data produced for PLC-3 and provided information on methodology and quality assurance for many laboratories.

In order to obtain relevant and reliable data in future stages of PLC, it is essential that the laboratories continue the implementation of the quality assurance programme to obtain international accredition. Establishing of working practices for quality assurance needs time. It is clear that the laboratories in the eastern Baltic countries still need support in terms of training and funding for improvement of analysis equipment.

Source Apportionment

In PLC-3 only the first steps were taken to distinguish between natural and anthropogenic contribution to riverine nutrient fluxes. Some of the Contracting Parties had not followed the Guidelines or had not provided the necessary information to carry out a proper source apportionment. Further, source apportionment has not been done at a comparable scale. Some of the Contracting Parties have carried out source apportionment on individual rivers, some on the sum of all monitored rivers, and some on the total load including load from unmonitored areas on a regional or national scale. The division between point sources and diffuse sources was problematic since load from freshwater fish farms and from stormwater constructions in some countries were included in the diffuse load, but in other countries in the point source load. Another problem was that some of the Contracting Parties could not deliver measured point source load figures, and that background load was not estimated by the same methodology. Apart from this, on some occasions source apportionment was carried out on gross riverine loads including retention and on other occasion only on net riverine loads. The former method gives a better estimate of the actual delivery from different load sources.

For PLC-4 it will be important to have a common definition of which sources constitute point sources and which constitute diffuse sources. A common methodology for estimating background (natural) load and retention is also required. Source apportionment should be carried out for all monitored rivers and for the Baltic Sea catchment areas of the Contracting Parties and for the nine Baltic Sea subregions. This means it will be necessary to develop a harmonised and comparable source apportionment approach. One of the most important prerequisites for that will be a point source inventory for the entire Baltic Sea catchment area. It should to some extent be possible to improve the source apportionment performed in PLC-3 to allow a comparison with a new source apportionment which will be carried out in PLC-4.

Data handling / data-entry system

Shortcomings in implementation of the PLC-3 Guidelines and in the reporting system with respect to time-tables and the data-entry system have caused many problems in data handling. The time-tables agreed by all Contracting Parties were not followed and were changed several times. Most of the data sets were incomplete and had very often to be corrected and amended by the countries concerned. Accordingly the agreed data outputs had to be changed several times during the PLC-3 process, too. In addition to that, many problems occurred with the dataentry system, developed as a Paradox application and built for storing of the

PLC-3 data. Due to the rapid development of computer technique this Paradox application did not fulfil the requirements of the Contracting Parties including the countries in transition. In the Paradox application the need to retrieve data stored in national databases into this programme was not foreseen, so that work was repeated in typing all data and information once more into the Paradox application. Furthermore, it was not possible to printout from the Paradox database to allow typing mistakes to be checked easily.

Although a data-entry system was implemented during PLC-3, there is a need to revise the existing database for future stages of PLC, to comply with the requirements of the Contracting Parties. It is necessary to minimise the work of the information providers and to minimise the risk of errors caused by modification and refeeding of the data. This will be only possible with a new data system for data transfer from the Contracting Parties to the data consultant, which will allow the Contracting Parties to copy data from their own databases into transfer files in the most convenient way. Such a new system of data transfer including an improved HELCOM waterborne pollution load database, should be taken into use in PLC-4.

The Contracting Parties only provided calculated annual load figures, so that checking of these figures by the data consultant was not possible. In future, whether or not primary data such as concentration and flow should be reported in addition to or instead of calculated loads must be taken into account, so that the data consultant has the opportunity to check and calculate load data on the basis of the agreed harmonised calculation procedures like in the database for the air deposition data.

7.3 Main conclusions and recommendations

The Third Baltic Sea Pollution Load Compilation (PLC-3) is a result of cooperation between all the Contracting Parties to the Helsinki Commission. So

far, three Baltic Sea Pollution Load Compilations have been carried out in 1987 (PLC-1), 1990 (PLC-2) and 1995 (PLC-3) with the aim of compiling the direct inputs of major pollutants entering the Baltic Sea from various sources (rivers, municipalities and industrial plants) on the basis of harmonised monitoring methods. Compared with the two former Pollution Load Compilations, PLC-3 is a significant step forward as it gives somewhat more reliable and complete data on the total loads on the Baltic Sea including the description of the methodologies used in the different countries. However, the quality of the reported load data during the three Pollution Load Compilations was very mixed, so that it is still impossible to make a realistic assessment of changes in the pollution load. There are many reasons for this, by which the most important ones being the following:

- PLC-1 was the first attempt to compile heterogeneous data that had been submitted to the Helsinki Commission on various occasions. Because the information came from various sources there were differences in the reliability and age of the data aswell as gaps in the data sets. Assuming that the values were often preliminary or based on very rough background information, it was recommended to use PLC-1 with caution.
- PLC-2 contained the generalised data characterising major pollution sources and loads with respect to $\mathrm{BOD}_{\scriptscriptstyle{7'}}\mathrm{N}_{\scriptscriptstyle\mathrm{total}}$ and $\mathrm{P}_{\scriptscriptstyle\mathrm{total}}.$ Information about untreated wastewater, overflows and bypasses from point sources discharging directly into the Baltic Sea, about the load coming from small rivers (< 5 m \sqrt{s}) and small settlements (< 10 000 PE) around the coast and coastal zones was missing. A quality assurance system was also missing.
- PLC-3 avoided the shortcomings in the coverage of the Baltic Sea catchment area, but the PLC-3 Guidelines were not fully implemented in all of the Contracting Party with respect to measuring

all the obligatory parameters in all pollution source categories. So, many uncertainties remain due to incomplete load data sets, especially from Russia, and from nearly all Contracting Parties with regard to heavy metals.

The assessment of the three Pollution Load Compilations has clearly shown that these could not be used for proving whether reduction targets (e.g. 50% reduction) have been met or not. Based on the data collected within the Pollution Load Compilations, proving that riverine pollution load represents the main pollution source, it is not possible to assess whether the goal of 50% reduction between 1987 and 1995 which was set as a target by the Ministerial Declaration in 1988 has been fulfilled. This is mainly due to the fact that riverine load data are highly dependent on meteorological factors such as precipitation and runoff and that the anthropogenic part could not be separated.

By ratifying the 1992 Convention, the Contracting Parties will implement relevant measures in the whole Baltic Sea catchment area in order to prevent pollution of the Baltic Sea. It is well known that the major part of the pollution load is transported by rivers to the Baltic Sea. The load of these rivers is caused by discharges from point and diffuse sources within the catchment areas of these rivers, and it is therefore an important task for the next PLCs to start investigations on collecting load data for point and diffuse sources situated within the whole Baltic Sea catchment area. This requires e.g. that data on point sources must be reported on a plant by plant basis in order to obtain information about the anthropogenic part of the river input and as a tool to determine whether the reduction goals have been met or not. Furthermore, data must be reported on diffuse sources, especially from agriculture. With these point and diffuse source inventories, together with a proper source apportionment, such goals as the 50% reduction of pollution load between 1987 and 1995 could be realistically evaluated.

List of References

Analytical Quality Assurance to Support the HELCOM PLC-3 Programme in the three Baltic States and Poland. Report to the EU PHARE Operational Services and HELCOM (Contract No 94-0891). Prepared by Dr. U. Lund, VKI and Dr. G. Topping, SOAEFD. Water Quality Institute (VKI), November 1996.

Arheimer, B., Brandt, M., Grahn, G., Roos E. and Sjöö, A. (1997): Modellerad kvävetransport, retention och källfördelning for södra Sverige. SMHI RH no. 13. Underlagsrapport till Naturvardsverkets uppdrag om Kväve fr n land till hav, 85 pp.

Behrendt, H. (1996): Inventories of point and diffuse sources and estimated nutrient loads - A comparison for different river basins in Central Europe. Water, Science & Technology, 33, 4-5, 99-107, 1996

HELCOM. Baltic Sea Environment Proceedings No. 45. Second Baltic Sea Pollution Load Compilation; 1993.

HELCOM. Baltic Sea Environment Proceedings No. 57. Guidelines for the Third Baltic Sea Pollution Load Compilation (PLC 3); 1994.

ISO/TR 13530, 1997. Water quality - Guide to analytical quality control for water samples.

ISO/IEC Guide 43-1, 1997. Proficiency Testing by Interlaboratory Comparisons - Development and Operation of Proficiency Testing Scheme.

Larson, U., Elmgren, R. and Wulff, F. Eutrophication and the Baltic Sea: Causes and consequences. AMBIO 14, 1985. pp. 9-14.

TC POLO 3/9 (Summary Record), Annex 5 "Guidelines for estimating natural and anthropogenic contributions to riverine fluxes (source apportionment)".

Annex 1/1

Annex 1. Analytical Methods

ORGANIC MATTER AND SUSPENDED SOLIDS

HEAVY METALS (Cd, Cr, Cu, Ni, Pb, Zn) AND MERCURY (Hg)

Annex 1/2

NUTRIENTS Analytical Principle Country parameter $P_{p_{n\alpha}}$ Molybdenum blue method DE, DK, EE, FI, LT, LV, PL, RU, SE P_{total} Digestion with peroxodisulphate; determination of orthophosphate DE, DK, EE, FI, LT, LV, PL, RU, SE by the molybdenum blue method P_{total} Digestion with peroxodisulphate + sulphuric acid; determination of FI (mainly wastewater) ortophosphate by the molybdenum blue method P_{total} Determination of orthophosphate by the molybdenum blue method; RU (some wastewater laboratories) estimation of P_{total} by correction factor N_{NHA} Indophenol blue method DE, DK, EE, FI, LV, PL, SE N_{N+4} Distillation and titration DK (a minority of laboratories, wastewater) N_{NHA} Distillation + Nessler method or titration Distance PL (wastewater) N_{N+4} Gas diffusion indicator method DK, FI (a minority of laboratories) N_{NH4} Nessler method LT, PL, RU N_{NLO3} Cadmium reduction method and determination of azo dye DE, DK, EE, FI, LT, LV, PL, RU (rivers), SE $N_{N_{NOS}}$ Devardas reduction DK (a minority of laboratories, wastewater) N_{NOS} Salicylate method Salicylate method EE (some wastewater laboratories), LV (wastewater), PL , RU (wastewater) N_{total} Peroxodisulphate digestion; reduction on cadmium column and DK, EE, FI, DE, LT, SE, RU determination of azo dye N_{total} Peroxodisulphate digestion; determination of nitrate by salicylate LT (wastewater), LV method N_{total} Kjeldahl plus determination of nitrate/nitrite **PL** N_{total} Reduction with Devarda's alloy; determination of ammonia by NK (a minority of laboratories), titration or indophenol blue titration or indophenol blue

Annex 2/1

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Annex 3/5

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Annex 4: Main information on municipal wastewater treatment plants in 1995

Annex 4/5

Annex 5: Main information on industrial plants in 1995

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Annex 6. Pollution Load Data in 1995

BALTIC SEA ENVIRONMENT PROCEEDINGS

- No. 1 JOINT ACTIVITIES OF THE BALTIC SEA STATES WITHIN THE FRAMEWORK OF THE CONVENTION ON THE PRO TECTION OF THE MARINE ENVIRONMENT OF THE BALTIC SEA AREA 1974-1978 $(1979)^*$
- No. 2 REPORT OF THE INTERIM COMMISSION (IC) TO THE BALTIC MARINE ENVIRONMENT PROTECTION **COMMISSION** (1981)*
- No. 3 ACTIVITIES OF THE COMMISSION 1980 -Report on the activities of the Baltic Marine Environment Protection Commission during 1980 -HELCOM Recommendations passed during 1980 $(1981)^*$
- No. 4 BALTIC MARINE ENVIRONMENT BIBLIOGRAPHY 1970-1979 (1981)*
- No. 5A ASSESSMENT OF THE EFFECTS OF POLLUTION ON THE NATURAL RESOURCES OF THE BALTIC SEA, 1980 PART A-1: OVERALL CONCLUSIONS $(1981)^*$
- No. 5B ASSESSMENT OF THE EFFECTS OF POLLUTION ON THE NATURAL RESOURCES OF THE BALTIC SEA, 1980 PART A-1: OVERALL CONCLUSIONS PART A-2: SUMMARY OF RESULTS PART B: SCIENTIFIC MATERIAL (1981)
- No. 6 WORKSHOP ON THE ANALYSIS OF HYDROCARBONS IN SEAWATER Institut für Meereskunde an der Universität Kiel, Department of Marine Chemistry, March 23 -April 3, 1981 (1982)
- No. 7A CTIVITIES OF THE COMMISSION 1981 -Report of the activities of the Baltic Marine Environment Protection Commission during 1981 including the Third Meeting of the Commission held in Helsinki 16-19 February 1982 -HELCOM Recommendations passed during 1981 and 1982 (1982)
- No. 8 ACTIVITIES OF THE COMMISSION 1982 -Report of the activities of the Baltic Marine Environment Protection Commission during 1982 including the Fourth Meeting of the Commission held in Helsinki 1-3 February 1983 -HELCOM Recommendations passed during 1982 and 1983 (1983)
- No. 9 SECOND BIOLOGICAL INTERCALIBRATION WORKSHOP Marine Pollution Laboratory and Marine Division of the National Agency of Environmental Protection, Denmark, August 17-20, 1982, Rønne, Denmark (1983)
- No. 10 TEN YEARS AFTER THE SIGNING OF THE HELSINKI CONVENTION National Statements by the Contracting Parties on the Achievements in Implementing the Goals of the Convention on the Protection of the Marine Environment of the Baltic Sea Area (1984)
- No. 11 STUDIES ON SHIP CASUALTIES IN THE BALTIC SEA 1979-1981 Helsinki University of Technology, Ship Hydrodynamics Laboratory, Otaniemi, Finland P. Tuovinen, V. Kostilainen and A. Hämäläinen (1984)

- No. 12 GUIDELINES FOR THE BALTIC MONITORING PROGRAMME FOR THE SECOND STAGE $(1984)^*$
- No. 13 ACTIVITIES OF THE COMMISSION 1983 -Report of the activities of the Baltic Marine Environment Protection Commission during 1983 including the Fifth Meeting of the Commission held in Helsinki 13-16 March 1984 -HELCOM Recommendations passed during 1983 and 1984 (1984)
- No. 14 SEMINAR ON REVIEW OF PROGRESS MADE IN WATER PROTECTION MEASURES 17-21 October 1983, Espoo, Finland (1985)
- No. 15 ACTIVITIES OF THE COMMISSION 1984 -Report of the activities of the Baltic Marine Environment Protection Commission during 1984 including the Sixth Meeting of the Commission held in Helsinki 12-15 March 1985 -HELCOM Recommendations passed during 1984 and 1985 (1985)
- No. 16 WATER BALANCE OF THE BALTIC SEA A Regional Cooperation Project of the Baltic Sea States; International Summary Report (1986)
- No. 17A FIRST PERIODIC ASSESSMENT OF THE STATE OF THE MARINE ENVIRONMENT OF THE BALTIC SEA AREA, 1980-1985; GENERAL CONCLUSIONS (1986)
- No. 17B FIRST PERIODIC ASSESSMENT OF THE STATE OF THE MARINE ENVIRONMENT OF THE BALTIC SEA AREA, 1980-1985; BACKGROUND DOCUMENT (1987)
- No. 18 ACTIVITIES OF THE COMMISSION 1985 -Report of the activities of the Baltic Marine Environment Protection Commission during 1985 including the Seventh Meeting of the Commission held in Helsinki 11-14 February 1986 -HELCOM Recommendations passed during 1986 $(1986)^*$
- No. 19 BALTIC SEA MONITORING SYMPOSIUM Tallinn, USSR, 10-15 March 1986 (1986)
- No. 20 FIRST BALTIC SEA POLLUTION LOAD COMPILATION (1987)
- No. 21 SEMINAR ON REGULATIONS CONTAINED IN ANNEX II OF MARPOL 73/78 AND REGULATION 5 OF AN NEX IV OF THE HELSINKI CONVENTION National Swedish Administration of Shipping and Navigation; 17-18 November 1986, Norrköping, Sweden (1987)
- No. 22 SEMINAR ON OIL POLLUTION QUESTIONS 19-20 November 1986, Norrköping, Sweden (1987)

- No. 23 ACTIVITIES OF THE COMMISSION 1986 -Report on the activities of the Baltic Marine Environment Protection Commission during 1986 including the Eighth Meeting of the Commission held in Helsinki 24-27 February 1987 -HELCOM Recommendations passed during 1987 $(1987)^*$
- No. 24 PROGRESS REPORTS ON CADMIUM, MERCURY, COPPER AND ZINC (1987)
- No. 25 SEMINAR ON WASTEWATER TREATMENT IN URBAN AREAS 7-9 September 1986, Visby, Sweden (1987)
- No. 26 ACTIVITIES OF THE COMMISSION 1987 -Report on the activities of the Baltic Marine Environment Protection Commission during 1987 including the Ninth Meeting of the Commission held in Helsinki 15-19 February 1988 -HELCOM Recommendations passed during 1988 (1988)
- No. 27A GUIDELINES FOR THE BALTIC MONITORING PROGRAMME FOR THE THIRD STAGE; PART A. INTRO DUCTORY CHAPTERS (1988)
- No. 27B GUIDELINES FOR THE BALTIC MONITORING PROGRAMME FOR THE THIRD STAGE; PART B. PHYSICAL AND CHEMICAL DETERMINANDS IN SEA WATER (1988)
- No. 27C GUIDELINES FOR THE BALTIC MONITORING PROGRAMME FOR THE THIRD STAGE; PART C. HARMFUL SUBSTANCES IN BIOTA AND SEDIMENTS (1988)
- No. 27D GUIDELINES FOR THE BALTIC MONITORING PROGRAMME FOR THE THIRD STAGE; PART D. BIOLOG ICAL DETERMINANDS (1988)
- No. 28 RECEPTION OF WASTES FROM SHIPS IN THE BALTIC SEA AREA - A MARPOL 73/78 SPECIAL AREA (1989)
- No. 29 ACTIVITIES OF THE COMMISSION 1988 -Report on the activities of the Baltic Marine Environment Protection Commission during 1988 including the Tenth Meeting of the Commission held in Helsinki 14-17 February 1989 -HELCOM Recommendations passed during 1989 (1989)
- No. 30 SECOND SEMINAR ON WASTEWATER TREATMENT IN URBAN AREAS 6-8 September 1987, Visby, Sweden (1989)
- No. 31 THREE YEARS OBSERVATIONS OF THE LEVELS OF SOME RADIONUCLIDES IN THE BALTIC SEA AFTER THE CHERNOBYL ACCIDENT Seminar on Radionuclides in the Baltic Sea 29 May 1989, Rostock-Warnemünde, German Democratic Republic (1989)
- No. 32 DEPOSITION OF AIRBORNE POLLUTANTS TO THE BALTIC SEA AREA 1983-1985 AND 1986 (1989)

- No. 33 ACTIVITIES OF THE COMMISSION 1989 -Report on the activities of the Baltic Marine Environment Protection Commission during 1989 including the Eleventh Meeting of the Commission held in Helsinki 13-16 February 1990 -HELCOM Recommendations passed during 1990 $(1990)^*$
- No. 34 STUDY OF THE RISK FOR ACCIDENTS AND THE RELATED ENVIRONMENTAL HAZARDS FROM THE TRANSPOR TATION OF CHEMICALS BY TANKERS IN THE BALTIC SEA AREA (1990)
- No. 35A SECOND PERIODIC ASSESSMENT OF THE STATE OF THE MARINE ENVIRONMENT OF THE BALTIC SEA, 1984-1988; GENERAL CONCLUSIONS (1990)
- No. 35B SECOND PERIODIC ASSESSMENT OF THE STATE OF THE MARINE ENVIRONMENT OF THE BALTIC SEA, 1984-1988; BACKGROUND DOCUMENT (1990)
- No. 36 SEMINAR ON NUTRIENTS REMOVAL FROM MUNICIPAL WASTE WATER 4-6 September 1989, Tampere, Finland (1990)
- No. 37 ACTIVITIES OF THE COMMISSION 1990 -Report on the activities of the Baltic Marine Environment Protection Commission during 1990 including the Twelfth Meeting of the Commission held in Helsinki 19-22 February 1991 -HELCOM Recommendations passed during 1991 (1991)
- No. 38 THIRD BIOLOGICAL INTERCALIBRATION WORKSHOP 27-31 August 1990, Visby, Sweden (1991)
- No. 39 AIRBORNE POLLUTION LOAD TO THE BALTIC SEA 1986-1990 (1991)
- No. 40 INTERIM REPORT ON THE STATE OF THE COASTAL WATERS OF THE BALTIC SEA (1991)
- No. 41 INTERCALIBRATIONS AND INTERCOMPARISONS OF MESUREMENT METHODS FOR AIRBORNE POLLUTANTS (1992)
- No. 42 ACTIVITIES OF THE COMMISSION 1991 -Report of the activities of the Baltic Marine Environment Protection Commission during 1991 including the 13th meeting of the Commission held in Helsinki 3-7 February 1992 -HELCOM Recommendations passed during 1992 (1992)
- No. 43 BALTIC MARINE ENVIRONMENT BIBLIOGRAPHY 1986-1990 (1992)
- No. 44 NITROGEN AND AGRICULTURE, INTERNATIONAL WORKSHOP 9-12 April 1991, Schleswig, Germany (1993)
- No. 45 SECOND BALTIC SEA POLLUTION LOAD COMPILATION (1993)
- No. 46 SUMMARIES OF THE PRE-FEASIBILITY STUDIES Prepared for the Baltic Sea Joint Comprehensive Environmental Action Programme (1993) *

- No. 47 HIGH LEVEL CONFERENCE ON RESOURCE MOBILIZATION Gdansk, Poland, 24-25 March 1993 Compilation of Presentations and Statements (1993)
- No. 48 THE BALTIC SEA JOINT COMPREHENSIVE ENVIRONMENTAL ACTION PROGRAMME (1993)
- No. 49 THE BALTIC SEA JOINT COMPREHENSIVE ENVIRONMENTAL ACTION PROGRAMME Opportunities and Constraints in Programme Implementation (1993)
- No. 50 SEMINAR ON RECEPTION FACILITIES IN PORTS Turku, Finland, 16-19 November 1992 (1993)
- No. 51 STUDY OF THE TRANSPORTATION OF PACKAGED DANGEROUS GOODS BY SEA IN THE BALTIC SEA AREA AND RELATED ENVIRONMENTAL HAZARDS (1993)
- No. 52 ACTIVITIES OF THE COMMISSION 1992 -Report on the activities of the Baltic Marine Environment Protection Commission during 1992 including the 14th meeting of the Commission held in Helsinki 2-5 February 1993 -HELCOM Recommendations passed during 1993 (1993)
- No. 53 BALTIC MARINE ENVIRONMENT BIBLIOGRAPHY 1991-1992 (1993)
- No. 54 FIRST ASSESSMENT OF THE STATE OF THE COASTAL WATERS OF THE BALTIC SEA (1993)
- No. 55 ACTIVITIES OF THE COMMISSION 1993 -Report on the activities of the Baltic Marine Environment Protection Commission during 1993 including the 15th meeting of the Commission held in Helsinki 8-11 March 1994 -HELCOM Recommendations passed during 1994 (1994)
- No. 56 INTERGOVERNMENTAL ACTIVITIES IN THE FRAMEWORK OF THE HELSINKI CONVENTION 1974-1994 (1994)
- No. 57 GUIDELINES FOR THE THIRD POLLUTION LOAD COMPILATION (PLC-3) (1994)
- No. 58 ICES/HELCOM WORKSHOP ON QUALITY ASSURANCE OF CHEMICAL ANALYTICAL PROCEDURES FOR THE BALTIC MONITORING PROGRAMME 5-8 October 1993, Hamburg, Germany (1994)
- No. 59 HELCOM SEMINAR FOR EXPERTS FROM ESTONIA, LATVIA, LITHUANIA AND RUSSIA ON THE IMPLEMENTA TION OF HELCOM ARRANGEMENTS, OTHER INTERNATIONAL INSTRUMENTS AND RELATED MATTERS 30 August - 3 September 1993, Riga, Latvia (1994)
- No. 60 ACTIVITIES OF THE COMMISSION 1994 -Report on the activities of the Baltic Marine Environment Protection Commission during 1994 including the 16th meeting of the Commission held in Helsinki 14-17 March 1995 -HELCOM Recommendations passed during 1995 (1995)

No. 61 RADIOACTIVITY IN THE BALTIC SEA 1984 - 1991 (1995)

- No. 62 ACTIVITIES OF THE COMMISSION 1995 -Report on the activities of the Baltic Marine Environment Protection Commission during 1995 including the 17th meeting of the Commission held in Helsinki 12-14 March 1996 -HELCOM Recommendations passed during 1996 (1996)
- No. 63 COASTAL AND MARINE PROTECTED AREAS IN THE BALTIC SEA REGION (1996)
- No. 64A THIRD PERIODIC ASSESSMENT OF THE STATE OF THE MARINE ENVIRONMENT OF THE BALTIC SEA, 1989-1993; EXECUTIVE SUMMARY (1996)
- No. 64B THIRD PERIODIC ASSESSMENT OF THE STATE OF THE MARINE ENVIRONMENT OF THE BALTIC SEA, 1989-1993; BACKGROUND DOCUMENT (1996)
- No. 65 OVERVIEW ON ACTIVITIES 1996 (1997)
- No. 66 BALTIC MARINE ENVIRONMENT BIBLIOGRAPHY 1993-1995 (1997)
- No. 67 WORKSHOP ON THE REDUCTION OF EMISSIONS FROM TRAFFIC IN THE BALTIC SEA AREA (1997)
- No. 68 THE EVALUATION OF THE RELATION OF ATMOSPHERIC DEPOSITION TO RIVERINE INPUT OF NITROGEN TO THE BALTIC SEA (1997) *
- No. 69 AIRBORNE POLLUTION LOAD TO THE BALTIC SEA 1991-1995 (1997)

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